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**Assessing the Potential Impacts of Declining Arctic Sea Ice Cover on the Photochemical  
Degradation of Dissolved Organic Matter in the Chukchi and Beaufort Seas**

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**Key Points:**

- Photochemical degradability of western Arctic Ocean DOM was tested in a solar simulator
- Solar irradiation caused significant decreases in CDOM light absorption at UV wavelengths
- Continued declines in sea ice cover should result in enhanced CDOM photodegradation

30   **Abstract**

31   A warming and shifting climate in the Arctic has led to significant declines in sea ice over the  
32   last several decades. Although these changes in sea ice cover are well documented, large  
33   uncertainties remain in how associated increases in solar radiation transmitted to the underlying  
34   ocean water column will impact heating, biological and biogeochemical processes in the Arctic  
35   Ocean. In this study, six under-ice marine, two ice-free marine, and two ice-free terrestrially-  
36   influenced water samples were irradiated using a solar simulator for 72 hours (representing ~10  
37   days of ambient sunlight) to investigate dissolved organic matter (DOM) dynamics from the  
38   Chukchi and Beaufort seas. Solar irradiation caused chromophoric DOM (CDOM) light  
39   absorption at 254 nm to decrease by 48 to 63%. An overall loss in total DOM fluorescence  
40   intensity was also observed at the end of all experiments, and each of 6 components identified by  
41   parallel factor analyses (PARAFAC) was shown to be photoreactive in at least one experiment.  
42   DOM fluorescence (FDOM) also indicated that the majority of DOM in under-ice and ice-free  
43   marine waters was likely algal-derived. Measurable changes in dissolved organic carbon (DOC)  
44   were only observed for sites influenced by riverine runoff. Losses of CDOM absorbance at  
45   shorter wavelengths suggest that the beneficial UV protection currently received by marine  
46   organisms may decline with the increased light transmittance associated with sea ice melt  
47   ponding and overall reductions of sea ice. Our FDOM analyses demonstrate that DOM  
48   irrespective of source was susceptible to photobleaching. Additionally, our findings suggest that  
49   photodegradation of CDOM in under-ice waters is not currently a significant source of carbon  
50   dioxide (CO<sub>2</sub>) (i.e., we did not observe systematic DOC loss). However, increases in primary  
51   production and terrestrial freshwater export expected under future climate change scenarios may  
52   cause an increase in CDOM quantity and shift in quality throughout Arctic Ocean surface waters.

53 As arctic temperatures continue to warm and summer sea ice further declines, examination of the  
54 resulting enhanced photodegradation processes and their impacts on the interplay between  
55 primary production, carbon cycling, and surface ocean heating processes will be paramount.  
56

57 **Index Terms and Keywords:** 0793, 0428, 0475, 0750, 1621; Chromophoric Dissolved Organic  
58 Matter (CDOM), Dissolved Organic Carbon (DOC), Light Absorption, Arctic, Sea Ice,  
59 Photodegradation  
60

## 61 **1. Introduction**

62 Recent climate warming has caused significant decreases in sea ice extent across the Arctic  
63 [Serreze *et al.*, 2007; Stroeve *et al.*, 2007; Comiso *et al.*, 2008], with a record minimum in  
64 September 2012 that was 49% below the 1979–2000 average [Perovich *et al.*, 2012]. Over the  
65 past three decades, there has also been an increase in the length of the summer melt season  
66 [Markus *et al.*, 2009] and an increasing percentage of thin first-year ice as compared with thicker  
67 multi-year ice [Maslanik *et al.*, 2007, Kwok and Rothrock, 2009; Maslanik *et al.*, 2011; Comiso,  
68 2012]. Declining sea ice extent reduces surface albedo, thereby enhancing the amount of solar  
69 radiation absorbed by associated ocean waters and ice. Furthermore, melt ponds form more  
70 easily and cover more surface area on thin first-year ice than on multi-year ice [Fetterer and  
71 Untersteiner, 1998; Nicolaus *et al.*, 2012], and light transmittance through these melt ponds can  
72 be up to an order of magnitude greater than through unpounded sea ice [Light *et al.*, 2008; Frey *et al.*,  
73 2011, Nicolaus *et al.*, 2012]. In future climate scenarios, Arctic sea ice is expected to  
74 continue thinning, decreasing in areal extent, and increasing its melt-pond coverage [Wang and  
75 Overland, 2009; Schröder *et al.*, 2014], yet large uncertainties remain in how the resulting

76 increase in solar radiation transmitted to the underlying ocean waters will impact biological and  
77 biogeochemical processes in the Arctic Ocean.

78

79 A major uncertainty linked to sea ice decline is to what extent dissolved organic matter (DOM)  
80 in underlying ocean waters will be impacted. In the Arctic Ocean, DOM primarily results from  
81 either: (a) *in situ* biological production in the upper water column and within sea ice; or (b)  
82 inputs of terrestrially-derived organic matter transported to the ocean by fluvial systems  
83 [Carlson, 2002; Holmes *et al.*, 2012; Raymond and Spencer, 2015]. Chromophoric DOM  
84 (CDOM) is the optically active fraction of this dissolved material, absorbing ultraviolet (UV) and  
85 visible light, and acting as one of the primary regulators of light penetration in surface waters  
86 [Blough and Del Vecchio, 2002; Nelson and Siegel, 2002]. CDOM regulates the amount of  
87 photosynthetically active radiation (PAR) in the water column, which influences the rates and  
88 distribution patterns of primary production [Retamal *et al.*, 2008]. CDOM also efficiently blocks  
89 UV radiation, thereby protecting marine organisms from DNA damage and other harmful effects  
90 [Williamson *et al.*, 2001]. By absorbing shortwave visible radiation (380–760 nm), CDOM can  
91 additionally contribute to the heating of surface waters and subsequent melting of sea ice [Kirk,  
92 1988, 1994; Pegau, 2002; Granskog *et al.*, 2007; Hill, 2008]. The absorption of light by CDOM  
93 can lead to photodegradation of DOM [Mopper *et al.*, 2015], generally resulting in a decrease in  
94 CDOM absorption (i.e., photobleaching; Helms *et al.*, 2008; 2014). Through these  
95 photodegradation processes, a pool of DOM can be directly photomineralized to inorganic  
96 carbon, with subsequent outgassing of CO<sub>2</sub> to the atmosphere [Stubbins *et al.*, 2008; Powers and  
97 Miller, 2015]. In addition, photochemical alteration of DOM influences the bioavailability of  
98 residual DOM to aquatic bacteria [Moran and Zepp, 1997; Bittar *et al.*, 2015]. The presence and

99 characteristics of arctic sea ice can therefore play a critical role in determining light penetration,  
100 in turn influencing CDOM distribution and biogeochemical cycling in surface ocean waters.  
101

102 Over the past ~30 years, the Chukchi and Beaufort seas of the western Arctic have experienced  
103 some of the greatest declines in sea ice extent across the Arctic Ocean [*Comiso et al.*, 2008],  
104 exposing the marine environment in this region to continually increasing amounts of solar  
105 radiation [*Perovich et al.*, 2007]. To date, only a few studies have investigated DOM in the  
106 western Arctic Ocean, with these typically focusing on the distribution of CDOM in open waters  
107 [*Gueguen et al.*, 2005, 2007, *Retamal et al.*, 2007; *Matsuoka et al.*, 2007, 2011; *Hill*, 2008;  
108 *Walker et al.*, 2009]. Furthermore, given the heavily riverine-influenced nature of the Arctic  
109 Ocean [*Opsahl et al.*, 1999; *Dittmar and Kattner*, 2003], DOM photoreactivity studies have  
110 focused primarily on terrigenous DOM in the coastal regions [*Belanger et al.*, 2006; *Osburn et*  
111 *al.*, 2009]. Thus, despite the critical need to understand the impacts of a shrinking and thinning  
112 sea ice cover, to the best of our knowledge this work represents the first study to focus upon the  
113 photoreactivity of DOM beneath sea ice. In this study, we used a series of controlled laboratory-  
114 based photodegradation experiments to investigate DOM dynamics in under-ice ocean waters  
115 collected from the Chukchi and Beaufort seas during the summer of 2011. Absorbance and  
116 fluorescence spectroscopy were used to identify the likely source of DOM, and to trace changes  
117 in CDOM optical properties during photochemical degradation. These results were then  
118 examined alongside changes in dissolved organic carbon (DOC) concentration to assess the  
119 relative importance of photobleaching and photomineralization. Finally, we discuss the potential  
120 photochemical susceptibility of Arctic Ocean DOM and how future changes to sea ice may

121 influence primary productivity, carbon biogeochemistry, and the heat balance in the Arctic  
122 Ocean.

## 124 **2. Methods**

### 125 **2.1. Sample Collection**

126 Sampling took place during the NASA Impacts of Climate Change on the Eco-Systems and  
127 Chemistry of the Arctic Pacific Environment (ICESCAPE) mission in the Chukchi and Beaufort  
128 seas during the summer melt season in July 2011 on the US Coast Guard Cutter (USCGC)  
129 *Healy*. Water samples were collected from 10 locations distributed across three north-south  
130 transects (Figure 1). The westernmost transect was located on the continental shelf of the  
131 Chukchi Sea (IFM<sub>1</sub> and ICM<sub>1-3</sub>). The central transect crossed the shelf-basin interface in the  
132 Chukchi Sea (IFM<sub>2</sub> and ICM<sub>4-5</sub>). The easternmost transect extended into a region of the Beaufort  
133 Sea basin impacted by river runoff, primarily from the Colville River (IFT<sub>1-2</sub> and ICM<sub>6</sub>).

134  
135 At six ice-covered stations (ICM<sub>1-6</sub>; Figure 1), water samples were collected from directly below  
136 the ice at the ice-water interface by hand-deploying a 2 L Kemmerer (vertically oriented) water  
137 sampler. At the remaining 4 stations located in open water, surface water samples were collected  
138 using the ship's CTD rosette, which consisted of a 12-place rosette with 30 L Niskin bottles.  
139 Two of these stations were located near the ice edge (IFM<sub>1-2</sub>; Figure 1) and two were from  
140 surface waters likely influenced by the Colville River (IFT<sub>1-2</sub>). Salinity measurements for each  
141 sample were made onboard using a Guildline Autosol salinometer. Immediately after sampling,  
142 water samples for photodegradation experiments were filtered through a pre-rinsed 0.2  $\mu$ m filter  
143 (Whatman PolyCap) and transferred into acid washed (10% HCl) precombusted (450°C for 6

144 hours) Kimax glass bottles. Samples were kept frozen (-20°C) and dark until return to the  
145 laboratory.

146

## 147 **2.2. Photodegradation Experiments**

148 After slowly thawing overnight, water samples were transferred into acid-washed (10% HCl),  
149 precombusted (450°C for 6 hours) round bottom quartz flasks (70 ml; Quartz Scientific Inc., OH,  
150 USA) and closed with ground-glass stoppers to exclude air bubbles. Flasks were rinsed with  
151 sample water three times prior to filling and then placed bottom side up in a 20°C water bath  
152 containing Milli-Q water. Samples were positioned just under the water surface. The bath was  
153 mounted inside an Atlas Suntest XLS+ solar simulator fitted with an Atlas NXe 1700w bulb and  
154 daylight filter, to provide a light spectrum similar to that of natural sunlight between 280 and 800  
155 nm. Total lamp power was  $765 \text{ W m}^{-2}$  with a constant exposure of  $65 \text{ W m}^{-2}$  between 300 to 400  
156 nm. Samples were exposed for 72 hours, receiving a radiant exposure of  $16,865 \pm 6.7 \text{ kJ m}^{-2}$  of  
157 energy (time integral of spectral irradiance), equivalent to approximately 10 days of ambient  
158 sunlight during the summer months at 70°N (based on the System for Transfer of Atmospheric  
159 Radiation (STAR) [Ruggaber *et al.*, 1994] and following adaptations described in Powers and  
160 Miller [2015]). CDOM, fluorescent DOM (FDOM) and DOC were measured (details below) at  
161 the beginning of the experiment and then again after 4, 12, 24, and 72 hours of light exposure.  
162 Triplicate samples and a single dark control (treated identically but wrapped in foil) were used at  
163 all time points, with triplicate dark controls for the initial and final (72-hour) time points.

164

165

166



### 167 2.3. CDOM Analysis

168 CDOM absorbance was measured using a Shimadzu UV-1800 UV-Visible spectrophotometer at  
169 1 nm intervals between 800 and 200 nm using a 10 cm quartz cuvette. All sample spectra were  
170 blank corrected and referenced against Milli-Q water (18.2 Ω). In order to minimize temperature  
171 effects, all measurements were made after samples had equilibrated to laboratory temperature.  
172 CDOM absorbance was assumed to be zero above 750 nm, therefore the average sample  
173 absorbance between 750 nm and 800 nm was subtracted from the spectrum to correct for offsets  
174 owing to instrument baseline drift, temperature, scattering and refractive effects [Green and  
175 Blough, 1994; Helms *et al*, 2008]. CDOM absorption coefficients were calculated from:

$$176 \quad a(\lambda) = 2.303A(\lambda)/l \quad (1)$$

178  
179 where  $a$  is the Napierian light absorption coefficient ( $\text{m}^{-1}$ ) at  $\lambda$  (the wavelength in nanometers),  $A$   
180 is the absorbance at  $\lambda$  wavelength, and  $l$  is the cell path length in meters [Green and Blough,  
181 1994].

182  
183 Spectral slopes ( $S$ ,  $\text{nm}^{-1}$ ) for each CDOM absorbance spectrum were derived using an  
184 exponential function:

$$185 \quad a(\lambda) = a(\lambda_0)e^{-S(\lambda-\lambda_0)} \quad (2)$$

187  
188 where  $a(\lambda)$  is the absorption coefficient of CDOM ( $\text{m}^{-1}$ ) at wavelength  $\lambda$  and  $\lambda_0$  is the reference  
189 wavelength. Spectral slopes were calculated across the wavelength ranges of 275–295 nm and

190 350–400 nm. Spectral slope parameters can provide information pertaining to DOM molecular  
191 weight, photochemical processing and source [Blough and Del Vecchio, 2002; Helms et al.,  
192 2008], and have been found to be largely independent of CDOM concentration [Brown, 1977].  
193 The 275–295 nm and 350–400 nm slope ranges were chosen as they have been shown to contain  
194 the greatest variation in a wide range of samples, including DOM-rich terrestrial, estuarine,  
195 coastal and highly photobleached waters [Helms et al., 2008; Spencer et al., 2012]. The ratio of  
196 these slopes ( $S_R$ ) has been shown to be particularly sensitive to changes in molecular weight and  
197 was calculated by dividing  $S_{275-295}$  by  $S_{350-400}$  [Helms et al., 2008]. All slopes are reported as  
198 positive numbers such that higher (i.e., steeper) slopes indicate a greater decrease in absorption  
199 with increasing wavelength.

200

#### 201 **2.4. FDOM Analysis**

202 The fluorescent portion of CDOM (FDOM) was characterized using fluorescence excitation  
203 emission matrix (EEM) spectroscopy, which creates three-dimensional structures composed of  
204 multiple emission spectra at a range of excitations and provides additional information about the  
205 chemical composition and sources of CDOM [Coble, 1996; 2007; Stedmon et al., 2003; Stubbins  
206 et al., 2014]. FDOM measurements were collected on a FluoroMax 4 spectrofluorometer (Horiba  
207 Jobin-Yvon). EEMs were obtained by recording sample emission across 320 to 500 nm (with 2  
208 nm increments) after excitation from 250 to 450 nm wavelengths (with 5 nm increments). A  
209 single sample EEM was collected at each time point (0, 4, 12, 24, and 72 hours). Owing to low  
210 fluorescence intensities, samples were run using long integration periods (0.5 s) to minimize  
211 measurement noise. All EEMs were blank corrected and Raman calibrated [Lawaetz and  
212 Stedmon, 2009; Murphy et al., 2010] using Milli-Q water spectra run on the same day. Finally, to

213 eliminate Rayleigh scattering effects, zeroes were inserted in the region where emission  
214 wavelengths are less than or equal to the excitation wavelength (+30 nm). The absorption  
215 coefficients of all samples were  $<10 \text{ m}^{-1}$ , eliminating the need for inner filter corrections  
216 [Stedmon and Bro, 2008; Murphy et al., 2010].

217

218 Many fluorescent components have overlapping peaks making it difficult to identify individual  
219 components within an EEM. To identify independently varying FDOM components within our  
220 dataset, and to assess changes in marine- vs. terrestrially-derived material, we conducted parallel  
221 factor analyses (PARAFAC) in MATLAB using the DOMFluor Toolbox following the methods  
222 as outlined in Stedmon and Bro [2008]. In total, we included 50 EEMs containing 5 different  
223 time points for each of the 10 experiments. By tracing processes using time-series measurements  
224 of FDOM, PARAFAC modeling may yield more meaningful results than if applied to discrete  
225 samples [Stedmon & Bro, 2008]. To simplify the modelling process and minimize noise, we  
226 included only emission wavelengths from 320 to 476 nm in the EEMs. No EEMs were identified  
227 as outliers. A series of PARAFAC models (utilizing between 2 to 7 components) were applied  
228 and initial model fitting assessed using the steps as outlined in Stedmon & Bro [2008]. Final  
229 model validation was conducted using a range of techniques including random split half analysis  
230 and random initialization [Stedmon & Bro, 2008]. Split half validation involves splitting the  
231 sample data into two halves and comparing model fits after running models on the two halves  
232 independently. Random initialization was then used to ensure that the models were in fact the  
233 least squares result and represented local minima.

234

235

## 236 **2.5. Dissolved Organic Carbon Analysis**

237 After the experiments, samples for DOC analyses were acidified ( $\text{pH} < 2$ ) by addition of HCl and  
238 analyzed for non-purgable organic carbon using a Shimadzu TOC-VCPH analyzer fitted with a  
239 Shimadzu ASI-V autosampler. Standards were prepared by the volumetric dilution of a stock  
240 solution containing 500  $\mu\text{M}$ -DOC (potassium hydrogen phthalate, p.a.) to produce the following  
241 series of standards: 0, 2, 5, 8, 10, 25, 50, 75, 100  $\mu\text{M}$ -DOC. In addition to standards, aliquots of  
242 deep seawater reference material (Batch 10, Lot# 05-10) from the Consensus Reference Material  
243 Project (CRM) were analyzed to ensure the precision and accuracy of the DOC analyses.  
244 Analyses of the CRM deviated by less than 5% from the reported value for these standards (41 to  
245 44  $\mu\text{M}$ -DOC; <http://yyy.rsmas.miami.edu/groups/biogeochem/Table1.htm>). Standard and sample  
246 volumes analyzed were 20 to 40 mL. Routine minimum detection limits in the investigators  
247 laboratory using the above configuration are  $2.8 \pm 0.3 \mu\text{M-C}$  and standard errors are typically  $1.7$   
248  $\pm 0.5$  % of the DOC concentration [Stubbins and Dittmar, 2012].

249

## 250 **3. Results**

### 251 **3.1. Photodegradation of CDOM**

252 Photodegradation experiments were conducted on waters from 10 separate stations across the  
253 Chukchi and Beaufort seas. Figure 2 shows absorption coefficient spectra over the course of a  
254 72-hour experiment (including triplicate analyses for each time step) for a typical ice-covered  
255 marine ( $\text{ICM}_1$ ), ice-free marine ( $\text{IFM}_1$ ), and ice-free terrestrially-influenced ( $\text{IFT}_1$ ) sample.  
256 CDOM measurements from replicate samples at each individual time step and experiment  
257 demonstrated little between-flask variability (with a coefficient of variance  $< 0.06$  at 254 nm for  
258 all experiments). Therefore, from here on we present only sample means of triplicate

measurements. All under-ice (ICM<sub>1-6</sub>) and marine samples (IFM<sub>1-2</sub>) had lower initial CDOM absorption values (ranging from 1.36 to 2.17 m<sup>-1</sup> at 254 nm) than terrestrially-influenced (IFT<sub>1-2</sub>) samples (3.12 to 5.79 m<sup>-1</sup> at 254 nm). Despite these initial differences, CDOM from all 10 stations was found to be photoreactive (Table 1, Figure 3a). For all experiments, photochemically induced changes in the absorption coefficient at 254 nm ( $a_{254}$ ) over the 72-hour period were found to be statistically significant ( $p < 0.05$ ; two-tailed Student's t-test on triplicate samples from the beginning and end of each experiment) with total losses in  $a_{254}$  of between 48 and 63%. No significant differences were observed between initial  $a_{254}$  values and those of the dark controls after 72 hours ( $p < 0.05$ ) across any of the experiments, confirming photochemical processes led to the observed reductions in absorbance. With the exception of samples ICM<sub>5</sub> and IFM<sub>1</sub>, sample waters also showed statistically significant photobleaching at 365 nm ( $a_{365}$ ), with a greater decrease observed in the IFT samples than in the marine samples (Table 2). Initial absorption coefficients at 440 nm ( $a_{440}$ ) were relatively low (most were at or below the detection limit of the spectrophotometer ( $\pm 0.002$  absorbance or 0.05 m<sup>-1</sup>)) and losses in  $a_{440}$  were not statistically significant for 7 of the 10 experiments (Table 2).

CDOM absorption values at 254 nm for the marine samples, whether from under-ice (ICM) or ice-free (IFM) waters, were not significantly different from one another (at both initial and final time points), yet both were statistically different ( $p < 0.05$ ) from terrestrially-influenced samples. Overall, the marine samples had lower initial  $a_{254}$  values (mean for ICM and IFM = 1.65 m<sup>-1</sup>) and exhibited smaller decreases in  $a_{254}$  (mean = 0.88 m<sup>-1</sup>) as compared to the terrestrially-influenced samples (mean  $a_{254}$  = 4.46 m<sup>-1</sup>, showing a mean loss of 2.63 m<sup>-1</sup>, Table 2).

As samples were not all optically thin, to more accurately determine changes in the efficiency of photobleaching between samples, the CDOM absorption coefficient was additionally plotted as a function of cumulative UV light (280–400 nm) absorbed (Figure 3b). The range of 280–400 nm was chosen because these are quantitatively the most important wavelengths for environmental photoreactions involving CDOM [Mopper *et al.*, 2015]. The decrease in the CDOM absorption coefficient at 254 nm was approximately exponential and could be modeled using the following three-parameter exponential decay equation:

$$C(u) = C_{\infty} + z_0 e^{-ku} \quad (3)$$

where  $u$  is cumulative UV light absorbed in kW/m<sup>2</sup>,  $C(u)$  is the modeled absorption coefficient,  $C_{\infty}$  is the non-photoreactive component,  $z_0$  is the photoreactive component, and  $k$  is the rate of decay. The exponential decay models fit the experimental results well with adjusted R<sup>2</sup> values ranging from ~0.94 to 0.98 for all experiments (Table 3). Overall, marine samples had smaller photoreactive components ( $z_0$ ) and residual nonphotoreactive ( $C_{\infty}$ ) components compared to the terrestrially influenced samples. Additionally, the rate of decay ( $k$ ) was greater for marine samples compared to the terrestrially-influenced samples, with the exception of IFM<sub>2</sub> which had a rate of decay similar to IFT<sub>2</sub>. There is also a strong relationship (adjusted R<sup>2</sup> = 0.99) between the initial absorption coefficient ( $C(0)$ ) and the photoreactive component ( $z_0$ ) (Figure 4).

Changes in DOM composition were also apparent during the 72-hour irradiation experiments with clear shifts in the spectral slope parameters. For all experiments,  $S_{275-295}$  decreased within the first 6 hours of irradiation, and then showed either little change, or a slight increase at

subsequent time points (Figure 5a). For most of the experiments,  $S_{350-400}$  decreased over time or showed little change (Figure 5b). Over the full length of the experiment, 9 of the 10 experiments (except IFT<sub>1</sub>) showed a statistically significant decrease ( $p < 0.5$ ; two-tailed Student's t-test on triplicate samples from the beginning and end of each experiment) in  $S_{275-295}$  (Table 4). For  $S_{350-400}$ , 9 of the 10 experiments (except ICM<sub>5</sub>) also showed a statistically significant ( $p < 0.5$ ) decrease (Table 4). Of the experiments that had a statistically significant change in  $S_R$ , all increased over the course of the experiment with the exception of ICM<sub>6</sub> (Table 4).

### 3.2. Photodegradation of FDOM and PARAFAC results

An overall loss in total DOM fluorescence intensity was observed at the end of all experiments. Figure 6 shows EEMs from four representative experiments (ICM<sub>1</sub>, IFM<sub>1</sub>, IFT<sub>1</sub>, IFT<sub>2</sub>) during the 72-hour irradiation period. Under-ice marine samples had consistently lower initial FDOM intensities as compared to ice-free marine and terrestrially-influenced samples. Furthermore, a greater proportion of the fluorescence signal was typically present at emission wavelengths less than 400 nm in the ICM and IFM samples, relative to the terrestrially-influenced sites. The two terrestrially-influenced samples clearly contained additional fluorescence signatures dominated by broad fluorescence peaks at emission wavelengths greater than 400 nm. Although all ten samples showed an overall decay in FDOM intensity over the 72-hour experiments, several samples showed an increase in total fluorescence intensity at intermediary time points (e.g. sample IFT<sub>2</sub> at t<sub>4</sub> in Figure 6), which then subsequently decayed.

Six independent fluorescent components were validated using PARAFAC analysis (Figure 7). Owing to the complex nature of DOM, it is unlikely that these components represent single

328 fluorophores, but instead represent a group of fluorophores that have similar properties and  
329 variability within the dataset (*Baker and Spencer, 2004; Murphy et al., 2010; Stubbins et al.,*  
330 *2014*). In particular, component 5 exhibited a shoulder that indicates that this component likely  
331 represents a mixture of fluorophores that the model was not able to separate because they co-  
332 varied in the dataset [*Stedmon and Markager, 2005a*]. The excitation and emission spectra of the  
333 components derived here are statistically compared to previously identified components in the  
334 OpenFluor database using the Tucker congruence coefficient in order to identify potential  
335 sources for each (*Murphy et al., 2014b*; Table 5). Components 1,3,4 and 5 were statistically  
336 similar to other components in the database and components 2 and 6 resembled components  
337 found in other studies (Table 5).

338  
339 Components 1, 2 and 6 had spectra resembling humic-like DOM, with broad emission spectra  
340 around and above 400 nm and a broader excitation spectra compared to the other components  
341 (Figure 7). Components 1 and 2 exhibit characteristics of humic-like terrestrially-derived  
342 material [*Stedmon and Markager, 2005a, b; Stedmon et al., 2007a; Kowalczyk et al., 2009;*  
343 *Yamashita et al., 2010b; Jørgensen et al., 2011; Murphy et al., 2011; Seredynska-Sobecka et al,*  
344 *2011; Shutova et al., 2014; Tanaka et al., 2014*]. Component 6 has also been identified as humic-  
345 like [*Stedmon and Markager, 2005a; Stedmon et al., 2007a; Stedmon et al., 2007b; Yamashita et*  
346 *al., 2010a; Jørgensen et al., 2011*] and *Stedmon and Markager [2005a]* categorized it as marine  
347 derived or anthropogenic in nature. Components 3, 4 and 5 exhibited spectra resembling  
348 ‘protein-like’ DOM, with emission spectra maxima below 400 nm. Component 3 most closely  
349 resembles amino acids, free or bound in proteins [*Murphy et al., 2008; 2014a*] and component 4  
350 resembles tyrosine [*Yamashita et al., 2011; Graeber et al., 2012; Yamashita et al., 2013*], both of



351 which are considered to be primarily microbially derived in marine waters [*Stedmon and*  
352 *Markager*, 2005b; *Coble*, 1996]. Component 5 is more difficult to classify, however, it is similar  
353 to component 6 in *Jørgensen et al.* [2011] that has been previously linked to surface water  
354 productivity.

355

356 Terrestrially-influenced samples IFT<sub>1</sub> and IFT<sub>2</sub> had the highest overall fluorescence at the  
357 beginning of the experiments and were largely dominated by the humic-like component 1 (29–  
358 40% of initial fluorescence; Table 6). The IFM<sub>1</sub> sample was initially dominated by component 5  
359 (48%) and also contained high levels of the protein-like components 3 (22%) and 4 (20%). The  
360 IFM<sub>2</sub> and under-ice samples generally exhibited low initial FDOM intensities with no clearly  
361 dominant fluorescent component. The exception was ICM<sub>6</sub>, which exhibited greater  
362 contributions from component 4 (56%).

363

364 Distinct similarities and differences between sample types were observed with respect to how  
365 components changed over the 72-hour irradiation experiments (Figure 8, Table 6). In general,  
366 under-ice marine waters exhibited only small changes in each component with irradiation  
367 (typically less than 0.03 r.u.). For the ice-free marine samples, IFM<sub>1</sub> exhibited large changes in  
368 fluorescence in the three protein-like components, whereas IFM<sub>2</sub> showed little change in any  
369 component. For the terrestrially-influenced samples, IFT<sub>1</sub> showed large losses in each  
370 component (55–99%) and IFT<sub>2</sub> exhibited large declines in components 2, 3, 4, and 5 (50–83%).  
371 Both IFT<sub>1</sub> and IFT<sub>2</sub> showed the greatest losses in ‘humic-like’ components 4 and 5 (80–99%).

372

373

### 374 3.3 Photomineralization of DOC

375 DOC concentrations from the beginning and end of nine of the experiments are shown in Table 1  
376 (sample loss prohibited measurement on the IFM<sub>1</sub> sample). Initial DOC concentrations in ICM  
377 and IFM samples were relatively low (53.6–71.9 µM), with higher concentrations in the two  
378 terrestrially-influenced samples (105.4 and 80.5 µM). There was no apparent relationship  
379 between salinity and DOC. Of the nine experiments, only three (ICM<sub>6</sub>, IFT<sub>1</sub>, and IFT<sub>2</sub>) showed a  
380 statistically significant ( $p < 0.05$ ) change in DOC at the end of the 72-hour irradiation period.  
381 Photomineralization losses were greatest in river-influenced samples, which showed an 8%  
382 (IFT<sub>1</sub>) and 7% (IFT<sub>2</sub>) decrease from initial DOC concentrations. The under-ice marine sample  
383 ICM<sub>6</sub> showed a 3% loss of DOC over this period. Overall losses in CDOM were far greater than  
384 for DOC (Figure 9a). In particular, the percent loss in  $a_{254}$  ranged from ~48 to 62%, whereas the  
385 percent loss in DOC was only ~0 to 8% (Figure 9b).

386

## 387 4. Discussion

### 388 4.1. Impact of Photodegradation on CDOM from Sea-Ice Melt

389 Initial CDOM absorption values were lower in ICM and IFM samples than in terrestrially-  
390 influenced samples, in good agreement with many previous studies that report typically higher  
391 CDOM absorption in coastal, river-influenced regions than in open ocean regions [Blough and  
392 Del Vecchio, 2002; Kitidis *et al.*, 2006; Osburn *et al.*, 2009]. CDOM in the water column has  
393 been shown to be excluded from sea ice during formation [Amon, 2003]. Therefore, in the  
394 absence of sea ice algae producing CDOM *in situ* [Scully and Miller, 2000], it would be expected  
395 that CDOM within sea ice melt would be low as well. The range in initial CDOM absorption  
396 values for ICM and IFM samples (1.36–2.17 m<sup>-1</sup>) could be partially explained by the range in

397 salinity (17.4–31.7), which suggests that DOM in these surface waters was of various sources  
398 and likely variable ages and residence times and therefore had undergone varying amounts of  
399 photochemical degradation prior to sampling. Despite differences in initial CDOM absorption  
400 values, CDOM was shown to be photoreactive irrespective of sampling location. Continuous  
401 losses of absorbance at 254 nm were observed (Figure 3a), suggesting photochemical exposure  
402 caused the average aromatic content of the DOM pool to be progressively reduced during  
403 irradiation [Osburn *et al.*, 2001; Stubbins *et al.*, 2010; Helms *et al.*, 2014].

404

405 Photobleaching kinetics for  $a_{254}$  in these experiments are well described by three parameter  
406 exponential decay models. Based on the photoreactive component ( $z_0$ ) predicted to be lost after  
407 infinite photo-exposure by these models,  $a_{254}$  could decrease by at least half for almost all of the  
408 samples if exposed to the sun indefinitely, indicating that UV absorption by CDOM at these  
409 shorter wavelengths is diminished by photodegradation. However, these models indicated that,  
410 regardless of sample type or initial CDOM values, residual amounts ( $C_\infty$ ) of  $a_{254}$  ( $\sim 0.57$ – $0.91 \text{ m}^{-1}$   
411 for under-ice and ice-free marine samples, and 1.36 and  $1.63 \text{ m}^{-1}$  for the two terrestrially  
412 influenced samples) would survive infinite photo-exposure, suggesting a persistent, yet small  
413 fraction of non-photoreactive DOM (that absorbs shortwave UV) was present in all waters.  
414 Additionally, there is also a strong relationship (adjusted  $R^2 = 0.99$ ) between the initial  
415 absorption coefficient ( $C(0)$ ) and the photoreactive component ( $z_0$ ) of all sample types (Figure 4)  
416 indicating that final  $a_{254}$  could potentially be predicted from initial measurements based on a  
417 regional model.

418

419 In our experiments, samples demonstrating significant changes in the spectral slope parameter,  
 420 generally exhibited decreases in both  $S_{275-295}$  and  $S_{350-400}$ , and an increase in  $S_R$  during  
 421 irradiation. For  $S_{275-295}$  in particular, much of the change occurred within the first 6 hours of the  
 422 72-hour irradiation (Figure 5a) indicating rapid initial photo-alteration in DOM composition.  
 423 Similar to our experiments, *Helms et al.* [2008] showed increases in  $S_R$  and decreases in  $S_{350-400}$ ,  
 424 however they showed that  $S_{275-295}$  increased during their photodegradation experiments. In  
 425 another study, *Spencer et al.* [2009] showed that (despite an initial increase in  $S_{275-295}$ )  $S_{275-295}$   
 426 values decreased in Congo River DOM during irradiation after extensive light exposure. We  
 427 suggest therefore that the decreases in  $S_{275-295}$  in most of our samples may have been caused by  
 428 extensive prior photo-degradation in the environment, although these optical patterns may also  
 429 derive from the freshly-produced algal nature of CDOM in sea ice impacted waters. For  
 430 example, irradiation of CDOM from algal cultures can result in  $S_{275-295}$  becoming shallower  
 431 during photobleaching [*Bittar et al.* 2015]. Changes in spectral slope parameters during  
 432 irradiation have been related to changes in mean molecular weight [*Helms et al.*, 2008], where  
 433 there have been particularly strong relationships between increases in  $S_R$  and decreases in  
 434 average DOM molecular weight. Five of our experiments (ICM<sub>1</sub>, ICM<sub>3</sub>, ICM<sub>4</sub>, IFM<sub>2</sub>, IFT<sub>1</sub>)  
 435 showed a significant increase in  $S_R$  during exposure. Based upon these results our samples may  
 436 have undergone a significant decrease in mean DOM molecular weight during irradiation.  
 437 However, the work by *Helms et al.* [2008] looked at samples across a strong terrigenous-to-  
 438 marine DOM gradient in the Southeastern US, far from the current study site in terms of  
 439 geography and biogeochemical system. Therefore, inferences about DOM quality based upon  
 440 spectral slope and  $S_R$  in our current system must be interpreted with due caution until work

441 shows that CDOM spectral qualities have similar empirical relationships to DOM quality in high  
442 latitude, sea ice impacted surface waters.

443

#### 444 **4.2. Impact of Photodegradation on FDOM from Sea-Ice Melt**

445 A greater relative proportion of fluorescence at emission wavelengths less than 400 nm was  
446 typically present in each of the under-ice and ice-free marine samples, as compared to the  
447 terrestrially-influenced sites. This is often referred to as ‘protein-like’ fluorescence, has  
448 commonly been reported in marine samples [Coble, 1996; Stedmon and Markager, 2001], and  
449 appears to be associated with a family of low molecular weight molecules, including aliphatics  
450 and nitrogen-enriched aromatics [Stubbins *et al.*, 2014]. The two terrestrially-influenced samples  
451 contained fluorescence signatures dominated by broad fluorescent peaks at emission wavelengths  
452 greater than 400 nm. This is typical of waters influenced by river runoff and commonly referred  
453 to as ‘humic-like’ fluorescence [Coble, 1996; Stedmon and Markager, 2001; Fellman *et al.*,  
454 2010]. These signatures have been associated with nitrogen-poor, relatively high mass molecular  
455 families [Stubbins *et al.*, 2014]. All ten samples show an overall decay in FDOM intensity over  
456 the 72-hour experiments, with the exception of protein-like production in several samples  
457 (occurring after 4 hours in sample IFT<sub>2</sub>, Figure 5), which subsequently decayed. Photo-induced  
458 production of tyrosine-like fluorescence (which contributes to the protein-like fluorescence  
459 signal) has been previously observed in lacustrine [Stedmon *et al.*, 2007a] and marine waters  
460 [Mann, 2010].

461

462 PARAFAC decomposition identified 6 components and overall, these fluorescent components  
463 were shown to decay, or remain constant, over the course of the experiments. Under-ice and ice–

464 free marine waters were dominated by the protein-like components, whereas terrestrially  
 465 influenced samples contained higher proportions of the humic-like fluorophores (Figure 8 and  
 466 Table 5). All components were shown to be photoreactive in at least one of our experiments,  
 467 further demonstrating that DOM irrespective of source was susceptible to photobleaching.  
 468 Under-ice and ice-free marine waters showed the greatest decreases in FDOM of protein-like  
 469 components (which are associated with recently algal-produced labile DOM), whereas the  
 470 greatest decreases in the terrestrially-influenced samples occurred in humic-like components.  
 471

472 Studies have indicated that photo-degradation of DOM can result in increasing [*Moran and*  
 473 *Zepp*, 1997] or decreasing biological lability [*Tranvik and Kokalj*, 1998] depending on DOM  
 474 source. In general, both field and laboratory studies suggest that photodegradation causes humic-  
 475 rich DOM to become more bioavailable whereas recently produced algal-derived DOM in  
 476 surface waters becomes less biolabile [*Benner and Biddanda*, 1998; *Obernosterer et al.*, 1999;  
 477 2001; *Tranvik et al.*, 1999; *Tranvik and Bertilsson*, 2001; *Bittar et al.*, 2015]. In particular,  
 478 protein-like fluorescence of algal origin has been shown to be both biolabile and photolabile,  
 479 indicating these two processes compete to mineralize fresh, algal DOC (*Stedmon and Markager*,  
 480 2005b; *Bittar et al.*, 2015). Based on our FDOM analyses, the majority of CDOM in under-ice  
 481 and ice-free marine waters appears to be algal-derived, therefore we may expect photoirradiation  
 482 of these waters to reduce the availability of biolabile DOC through direct photo-mineralization  
 483 and through photo-alteration to less biolabile material as found for DOM produced by algal  
 484 cultures [*Bittar et al.*, 2015]. Thus, photochemistry of algal-derived DOM in sea ice systems may  
 485 short-circuit the microbial loop, whereas photo-degradation of terrigenous DOM is likely to  
 486 increase DOM biolability and enhance bacterioplankton productivity.

#### 487     **4.3. Photobleaching vs. Photomineralization**

488     Three of our samples displayed significant changes in DOC concentration with irradiation. The  
489     three samples (ICM<sub>6</sub>, IFT<sub>1</sub> and IFT<sub>2</sub>) were all from the easternmost transect that was closest to  
490     the mouth of the Colville River and were likely influenced by the relatively high-DOC, high-  
491     CDOM riverine water runoff. Investigating the various components identified in the PARAFAC  
492     analysis, changes in component 2 (identified as humic-like and terrestrially-derived; Table 5) had  
493     the strongest significant correlation with changes in DOC ( $R = 0.78$ ,  $p < 0.02$ ) indicating that  
494     there may be a link between this component and DOC. These results suggest that as melt ponds  
495     form and sea ice retreats, terrestrial DOM below the ice would be susceptible to  
496     photomineralization to dissolved inorganic carbon.

497  
498     Our experiments suggest that it is not possible to calculate DOC from CDOM using a constant,  
499     linear relationship in the Chukchi and Beaufort seas. Several methods for using CDOM  
500     measurements as a proxy for DOC have previously been proposed [*Ferrari et al.*, 1996; *Del*  
501     *Vecchio and Blough*, 2004; *Guéguen et al.*, 2005] because CDOM can be measured at a much  
502     higher spatial resolution using remote sensing and at a high temporal resolution using *in situ*  
503     measurements, and as such, it has the potential to more easily provide broader spatial and  
504     temporal coverage. A promising new approach was suggested by *Fichot and Benner* [2011],  
505     which derives DOC concentrations from CDOM absorption coefficients at 275 and 295 nm and  
506     their approach worked well for the coastal waters of the Beaufort Sea (eastward of our study  
507     area). However, based on the results of our photodegradation experiments, the method is likely  
508     not applicable for the nearby under-ice and ice-free marine waters (that are less influenced by

509 river runoff) as while there were measurable changes in the absorption coefficients at shorter  
510 wavelengths with irradiation, there were no measurable changes in DOC.  
511

512 Photodegradation resulted in measurable changes in the optical properties of DOM in all sample  
513 waters, yet none of the ice-free and under-ice marine samples from the westernmost and central  
514 transects showed a significant change in DOC with irradiation. Several studies have shown that  
515 photochemical loss of DOC is less efficient than CDOM photobleaching [*Moran et al.*, 2000;  
516 *Vahatalo and Wetzel*, 2004; *Spencer et al.*, 2009]. Furthermore, *Osburn et al.* [2009] conducted  
517 photoreactivity experiments on water samples from the Mackenzie River, estuary, shelf and gulf  
518 regions of the western Canadian Arctic and found less than 1% loss in DOC after 3 days in  
519 ambient arctic sunlight. Similar results were found for sea ice brine from Antarctic sea ice  
520 [*Norman et al.*, 2011]. The absence of measurable changes in DOC in waters in the westernmost  
521 and central transects in this study indicates that either there is a decoupling between CDOM  
522 photobleaching and DOC photomineralization or that any changes in DOC were below the  
523 detection limit of our instrument.  
524

525 Photochemistry can result in the production of CO<sub>2</sub> through two different pathways: (a) the  
526 abiotic photomineralization of DOC into CO<sub>2</sub> and CO; and (b) a photochemical-biotic pathway  
527 through which photodegradation alters the bioavailability of DOM potentially making it more  
528 labile allowing it to be consumed by bacteria and converted into biomass and CO<sub>2</sub> [see sources in  
529 *Mopper et al.*, 2015]. Not only was there very little photomineralization evident in the western  
530 and central transects in our study, but the changes in spectral slopes and fluorescence  
531 components observed in these experiments suggest that there may have been an overall reduction



532 in the biolability of DOM in under-ice and marine surface waters. Therefore, it appears that  
533 photodegradation of marine waters and sea ice melt does not result in significant outgassing of  
534 CO<sub>2</sub> to the atmosphere via either the purely photochemical pathway or the combined  
535 photochemical-biological pathway. Results presented here suggest that photodegradation of  
536 under-ice waters is far more important for photobleaching (and changes in light penetration in  
537 the euphotic zone) than it is for carbon cycling.

538

## 539 **5. Implications and Conclusions**

540 As the Arctic continues to warm and sea ice extent declines, the photoreactivity of DOM may  
541 have important implications for primary productivity and the heat balance of the Arctic Ocean.  
542 Thirty years ago much of the Arctic Ocean was covered with thick multi-year ice throughout the  
543 summer [Serreze *et al.*, 2007; Kwok and Rothrock, 2009], which limited the amount of UV and  
544 visible light reaching under-ice waters. With decreasing sea ice extent and increases in thin first-  
545 year ice and surface melt ponds, the surface layer of the Arctic Ocean is being exposed to more  
546 solar radiation during the summer melt season. Several studies suggest that primary productivity  
547 will increase in arctic waters owing to this decrease in sea ice extent [Arrigo *et al.*, 2008; Arrigo  
548 and van Dijken, 2011] and increase in melt pond coverage [Arrigo *et al.*, 2012] and the  
549 corresponding increase in PAR reaching ocean surface waters. However, continued ponding,  
550 thinning, and ultimate loss of summer sea ice cover would also lead to a decrease in UV  
551 absorption by surface CDOM with photodegradation. Combined with the unprecedented loss in  
552 Arctic ozone in recent years [Solomon *et al.*, 2007; Manney *et al.*, 2011], this may result in  
553 exposure of marine organisms to more harmful UV radiation and may in turn negatively impact  
554 primary productivity. An additional complication is that the decrease in sea ice extent and

555 increase in open water during the summer months have been shown to coincide with an increase  
556 in cloud coverage [Wang and Key, 2005; Eastman and Warren, 2010; Palm *et al.*, 2010], which  
557 limits the amount of shortwave radiation reaching the ocean surface. Models show that these  
558 trends will likely continue in the future [Vavrus *et al.*, 2010] potentially counteracting the  
559 increase in UV exposure owing to changes in sea ice, but also limiting the amount of available  
560 PAR [Belanger *et al.*, 2013]. This presents challenges for predicting future trends in the amount  
561 of UV and visible radiation reaching surface waters and quantifying the overall impact that  
562 CDOM photodegradation will have on primary production.

563

564 Warming temperatures in the Arctic are leading to permafrost thaw [Romanovsky *et al.*, 2010]  
565 and increases in riverine discharge [Peterson *et al.*, 2002], which in turn will likely lead to  
566 increases in terrestrial DOM output to surface waters of the Arctic Ocean [Frey and McClelland,  
567 2009]. Our results suggest that increasing terrestrial DOM supply will result in a measureable  
568 increase in DOM photodegradation and loss of DOC, since significant changes in DOC were  
569 observed in samples exhibiting greater contributions from humic-like terrestrially sourced  
570 FDOM components. Increases in daily minimum flow rates, particularly in winter, have been  
571 shown for arctic rivers [Smith *et al.*, 2007], which could potentially increase the amount of  
572 terrestrial DOM transported to surface waters of the Arctic Ocean [Holmes *et al.*, 2008; Stedmon  
573 *et al.*, 2011]. Since this would occur when there is little light north of the Arctic Circle and the  
574 Arctic Ocean is largely covered in sea ice, this DOM would initially be protected from  
575 photodegradation and could be transported well into shelf seas and possibly the Arctic Ocean  
576 basin interior. Furthermore, the spring freshet period has been shown to export a major fraction  
577 of the annual DOM load of Arctic rivers and DOM exported during this period is relatively more

578 aromatic in character and more photolabile in comparison to other times of year [*Spencer et al.*,  
579 2008; *Holmes et al.*, 2012; *Mann et al.*, 2012]. This major pulse of photoreactive DOM enters the  
580 Arctic Ocean at the onset of spring and thus with the decrease in sea ice extent and increase in  
581 melt ponding during the following spring and summer, our results suggest that this DOM would  
582 be poised for significant photodegradation during the subsequent melt season. This combined  
583 with increased DOM from under-ice blooms and subsequent photodegradation could alter CO<sub>2</sub>  
584 saturation in surface waters.

585  
586 By absorbing sunlight and subsequently re-emitting it as heat, CDOM also contributes to the  
587 heating of surface waters [*Kirk*, 1988]. With increasing sea ice melt, more light may be  
588 transmitted through sea ice and absorbed and emitted as heat by the existing CDOM below. This  
589 increase in heat absorption by the under-ice water column could increase local sea surface  
590 temperatures and potentially lead to further sea ice melt from below, thus initiating a positive  
591 feedback to sea ice melt [*Hill*, 2008]. A future increase in CDOM supply, particularly from  
592 riverine sources, would considerably increase CDOM absorption and it is likely that  
593 photodegradation will substantially decrease the light absorption of this additional riverine  
594 CDOM. At present, the combined effect of increases in CDOM additions from *in situ*  
595 production and riverine sources as well as increasing CDOM photobleaching as ice cover  
596 recedes upon ocean surface heating remain unclear.

597  
598 In conclusion, our observed reductions in CDOM absorbance at shorter wavelengths suggest that  
599 the beneficial UV protection currently received by marine organisms may significantly reduce  
600 with sea ice melt ponding and overall declines of sea ice cover. However, some of these CDOM

601 losses may be balanced by increasing CDOM inputs from increased primary production and  
602 terrestrial carbon export. We additionally show that under current conditions, photodegradation  
603 of CDOM in under-ice waters is not a significant source of CO<sub>2</sub>, and it is unclear whether the  
604 potential for surface ocean heating is compromised with irradiation. However, as arctic  
605 temperatures warm and summer sea ice continues to disappear, continued examination of the  
606 resulting enhanced photodegradation processes and their impacts on the interplay between  
607 primary production, carbon biogeochemistry, and surface ocean heating will be vital.

608

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## References

- Amon, R.M.W. (2003) The role of dissolved organic matter for the organic carbon cycle in the Arctic Ocean, in *The Organic Carbon Cycle in the Arctic Ocean*, edited by R. Stein R. MacDonald, pp. 83-99, Springer, Berlin.
- Arrigo, K. R. and G. L. van Dijken (2011), Secular trends in Arctic Ocean net primary production, *Journal of Geophysical Research*, 116, C09011, doi:10.1029/2011JC007151.
- Arrigo, K. R., G. van Dijken and S. Pabi (2008), Impact of a shrinking Arctic ice cover on marine primary production, *Geophysical Research Letters*, 35, L19603, doi:10.1029/2008GL035028.
- Arrigo, K. R. et al. (2012), Massive phytoplankton blooms under Arctic sea ice, *Science*, 336, 1407-1408.
- Baker, A., and R. G. M. Spencer, (2004), Characterization of dissolved organic matter from source to sea using fluorescence and absorbance spectroscopy, *Science of the Total Environment* 333, 217-232.
- Belanger, S., H. Xie, N. Krotkov, and P. Larouche (2006) Photomineralization of terrigenous dissolved organic matter in Arctic coastal waters from 1979 to 2003: Interannual variability and implications of climate change, *Global Biogeochemical Cycles*, 20, GB4005, doi:10.1029/2006GB002708.
- Belanger, S., M. Babin and J.-E. Tremblay (2013), Increasing cloudiness in Arctic damps the increase in phytoplankton primary production due to sea ice receding, *Biogeosciences*, 10, 4087–4101, doi:10.5194/bg-10-4087-2013.
- Benner, R., and B. Biddanda (1998), Photochemical transformations of surface and deep marine dissolved organic matter: Effects on bacterial growth, *Limnol. Oceanogr.*, 43(6), 1373-1378.
- Bittar, T. B., A. A. H. Vieira, A. Stubbins and K. Mopper, (2015), Competition between photochemical and biological degradation of dissolved organic matter from the cyanobacteria *Microcystis aeruginosa*, *Limnology and Oceanography*, 60(4), 1172-1194.
- Brown, M. (1977), Transmission spectroscopy examinations of natural waters, *Estuar. Coast. Mar. Sci.*, 5, 309-317, doi: 10.1016/0302-3524(77)90058-5.
- Blough, N.V., and R. Del Vecchio (2002), Chromophoric DOM in the coastal environment, in *Biogeochemistry of Marine Dissolved Organic Matter*, edited by D.A. Hansell and C.A. Carlson, pp. 509-546, Elsevier, San Diego, Calif.

668 Carlson (2002), Production and Removal Processes, in Biogeochemistry of Marine Dissolved  
 669 Organic Matter, edited by D.A. Hansell and C.A. Carlson, pp. 91-151, Elsevier, San Diego,  
 670 Calif.  
 671  
 672 Coble, P.G. (1996). Characterization of marine and terrestrial DOM in seawater using excitation-  
 673 emission matrix spectroscopy, *Mar. Chem.*, 51, 325-346.  
 674  
 675 Coble, P.G. (2007), Marine Optical Biogeochemistry: The Chemistry of Ocean Color, *Chem.*  
 676 *Rev.*, 107, 402-418.  
 677  
 678 Comiso, J. C., C. L. Parkinson, R. Gersten, and L. Stock (2008), Accelerated decline in the  
 679 Arctic sea ice cover, *Geophys. Res. Lett.*, 35, L01703, doi:10.1029/2007GL031972.  
 680  
 681 Comiso, J. C. (2012), Large decadal decline of the Arctic multiyear ice cover, *J. Clim.*, 25(4),  
 682 1176–1193, doi:10.1175/JCLI-D-11-00113.1.  
 683  
 684 Del Vecchio, R., and N. V. Blough (2004), Spatial and seasonal distribution of chromophoric  
 685 dissolved organic matter and dissolved organic carbon in the Middle Atlantic Bight, *Mar. Chem.*,  
 686 89(1–4), 169–187, doi:10.1016/j.marchem.2004.02.027.  
 687  
 688 Dittmar, T., and G. Kattner (2003), The biogeochemistry of the river and shelf ecosystem of the  
 689 Arctic Ocean: a review, *Marine Chemistry*, 83, 103-120, doi:10.1016/S0304-4203(03)00105-1.  
 690  
 691 Eastman, R. and S. G. Warren (2010) Arctic Cloud Changes from Surface and Satellite  
 692 Observations, *Journal of Climate*, 23, 4233-4242, doi:10.1175/2010JCLI3544.1.  
 693  
 694 Fellman, J.R., E. Hood, R.G.M. Spencer, (2010), Fluorescence spectroscopy opens new windows  
 695 into dissolved organic matter dynamics in freshwater ecosystems: A review, *Limnology and*  
 696 *Oceanography*, 55, 2452-2462.  
 697  
 698 Ferrari, G. M., M. D. Dowell, S. Grossi, and C. Targa (1996), Relationship between the optical  
 699 properties of chromophoric dissolved organic matter and total concentration of dissolved organic  
 700 carbon in the southern Baltic Sea region, *Mar. Chem.*, 55(3–4), 299–316.  
 701  
 702 Fetterer, F. and N. Untersteiner (1998), Observations of melt ponds on Arctic sea ice, *Journal of*  
 703 *Geophysical Research*, 103(C11), 24, 821–24,835.  
 704  
 705 Fichot, C. G. and R. Benner (2011), A novel method to estimate DOC concentrations from  
 706 CDOM absorption coefficients in coastal waters, *Geophysical Research Letters*, 38, L03610,  
 707 doi:10.1029/2010GL046152.  
 708  
 709 Frey, K. E. and J. W. McClelland (2009), Impacts of permafrost degradation on arctic river  
 710 biogeochemistry, *Hydrological Processes*, 23, 169-182, doi:10.1002/hyp.7196.  
 711  
 712 Frey, K. E., D. K. Perovich, and B. Light (2011), The spatial distribution of solar radiation under  
 713 a melting Arctic sea ice cover, *Geophys. Res. Lett.*, 38, L22501, doi:10.1029/2011GL049421.

714  
715 Graeber, D., J. Gelbrecht, M. Pusch, C. Anlanger, and D. von Schiller (2012), Agriculture has  
716 changed the amount and composition of dissolved organic matter in Central European headwater  
717 streams, *Science of The Total Environment*, 438, 435-446.  
718  
719 Granskog, M. A., R. W. Macdonald, C. J. Mundy and D. G. Barber (2007) Distribution,  
720 characteristics and potential impacts of chromophoric dissolved organic matter (CDOM) in  
721 Hudson Strait and Hudson Bay, Canada, *Continental Shelf Research*, 27, 2032-2050.  
722 doi:10.1016/j.csr.2007.05.001.  
723  
724 Green, S. A., and N. V. Blough (1994), Optical absorption and fluorescence properties of  
725 chromophoric dissolved organic matter in natural waters, *Limnol. Oceanogr.*, 39, 1903-1916.  
726  
727 Gueguen, C., L. Guo, and N. Tanaka (2005), Distributions and characteristics of colored dissolved  
728 organic matter in the Western Arctic Ocean, *Continental Shelf Research*, 25, 1195-1207,  
729 doi:10.1016/j.csr.2005.01.005.  
730  
731 Gueguen, C., L., Guo, M. Yamamoto-Kawai, and N. Tanaka (2007), Colored dissolved organic  
732 matter dynamics across the shelf-basin interface in the western Arctic Ocean, *Journal of*  
733 *Geophysical Research*, 112, C05038, doi:10.1029/2006JC003584.  
734  
735 Helms, J. R., A. Stubbins, J. D. Ritchie, E. C. Minor, D. J. Kieber, and K. Mopper (2008),  
736 Absorption spectral slopes, and slope ratios as indicators of molecular weight, source, and  
737 photobleaching of chromophoric dissolved organic matter, *Limnol. Oceanogr.*, 53, 955-969.  
738  
739 Helms, J.R., J. Mao, A. Stubbins, K. Schmidt-Rohr, R.G.M. Spencer, P.J. Hernes and K. Mopper  
740 (2014), Loss of optical and molecular indicators of terrigenous dissolved organic matter during  
741 long-term photobleaching. *Aquatic Sciences.*, 76(3), 353-373.  
742  
743 Hill, V. J. (2008), Impacts of chromophoric dissolved organic material on surface ocean heating  
744 in the Chukchi Sea, *Journal of Geophysical Research*, 113, C07024, doi:10.1029/2007JC004119.  
745  
746 Holmes, R. M., J. W. McClelland, P. A. Raymond, B. B. Frazer, B. J. Peterson and M. Stieglitz  
747 (2008), Lability of DOC transported by Alaskan rivers to the Arctic Ocean, *Geophysical*  
748 *Research Letters*, 35, L03402, doi:10.1029/2007GL032837.  
749  
750 Holmes, R. M., J. W. McClelland, B. J. Peterson, S. E. Tank, E. Bulygina, T. I. Eglinton, V. V.  
751 Gordeev, T. Y. Gurtovaya, P. A. Raymond, D. J. Repeta, R. Staples, R. G. Striegl, A. V.  
752 Zhulidov, and S. A. Zimov (2012), Seasonal and annual fluxes of nutrients and organic matter  
753 from large rivers to the Arctic Ocean and surrounding seas, *Estuaries and Coasts* 35, 369-382,  
754 doi: 10.1007/s12237-011-9386-6.  
755  
756 Jørgensen, L., C. A. Stedmon, T. Kragh, S. Markager, M. Middelboe, and M. Søndergaard  
757 (2011), Global trends in the fluorescence characteristics and distribution of marine dissolved  
758 organic matter, *Mar. Chem.*, 126, 139-148, doi:10.1016/j.marchem.2011.05.002.  
759

760 Kirk, J.T.O. (1988), Solar heating of water bodies as influenced by their inherent optical  
 761 properties, *Journal of Geophysical Research*, 93, D9, 10897–10908.  
 762  
 763 Kirk, J. T. O. (1994), *Light and Photosynthesis in Aquatic Ecosystems*, Cambridge Univ. Press,  
 764 Cambridge, U.K.  
 765  
 766 Kitidis, V., A. P. Stubbins, G. Uher, R. C. U. Goddard, C. S. Law and E. M. S. Woodward  
 767 (2006), Variability of chromophoric organic matter in surface waters of the Atlantic Ocean, *Deep*  
 768 *Sea Research Part II: Topical Studies in Oceanography*, 53(14), 1666-1684.  
 769  
 770 Kowalczuk, P., M. J. Durako, H. Young, A. E. Kahn, W. J. Cooper, and M. Gonsior (2009),  
 771 Characterization of dissolved organic matter fluorescence in the South Atlantic Bight with use of  
 772 PARAFAC model: Interannual variability, *Marine Chemistry*, 113, 182-196.  
 773  
 774 Kwok, R., and D. A. Rothrock (2009), Decline in Arctic sea ice thickness from submarine and  
 775 ICESat records: 1958-2008, *Geophys. Res. Lett.*, 36, L15501, doi:10.1029/2009GL039035.  
 776  
 777 Lawaetz, A.J., and C.A. Stedmon (2009), Fluorescence intensity calibration using the Raman  
 778 scatter peak of water, *Appl. Spectrosc.*, 63 (8), 936-940.  
 779  
 780 Light, B., T. C. Grenfell, and D. K. Perovich (2008), Transmission and absorption of solar  
 781 radiation by Arctic sea ice during the melt season, *J. Geophys. Res.*, 113, C03023,  
 782 doi:10.1029/2006JC003977.  
 783  
 784 Mann, P. J. (2010) *Ammonium Photoproduction and CDOM Biogeochemistry in Marine*  
 785 *Systems*. PhD thesis. Newcastle University, UK.  
 786  
 787 Mann, P.J., A. Davidova, N. Zimov, R.G.M. Spencer, S. Davydov, E. Bulygina, S. Zimov, R.M.  
 788 Holmes (2012), Controls on the composition and lability of dissolved organic matter in Siberia's  
 789 Kolyma River Basin, *J. Geophys. Res.* 117, G01028, doi:10.1029/2011JG001798.  
 790  
 791 Manney, G. L. et al. (2011), Unprecedented Arctic ozone loss in 2011, *Nature*, 478, 469-475,  
 792 doi:10.1038/nature10556.  
 793  
 794 Markus, T., J. C. Stroeve, and J. Miller (2009), Recent changes in Arctic sea ice melt onset,  
 795 freezeup, and melt season length, *J. Geophys. Res.*, 114, C12024, doi:10.1029/2009JC005436.  
 796  
 797 Maslanik, J.A., C. Fowler, J. Stroeve, S. drobot, J. Zwally, D. Yi, and W. Emery (2007), A  
 798 younger, thinner Arctic ice cover: Increased potential for rapid, extensive sea-ice loss, *Geophys.*  
 799 *Res. Lett.*, 34, L24501, doi:10.1029/2007GL032043.  
 800  
 801 Maslanik, J., J. Stroeve, C. Fowler, and W. Emery (2011), Distribution and trends in Arctic sea  
 802 ice age through spring 2011, *Geophys. Res. Lett.*, 38, L13502, doi:10.1029/2011GL047735.  
 803



804 Matsuoka, A., Y. Huot, K. Shimada, S. Saitoh and M. Babin (2007), Bio-optical characteristics  
805 of the western Arctic Ocean: implications for ocean color algorithms, *Can. J. Remote Sensing*,  
806 33(6), 503-518, doi  
807  
808 Matsuoka, A., V. Hill, Y. Huot, M. Babin and A. Bricaud (2011), Seasonal variability in the light  
809 absorption properties of western Arctic waters: Parameterization of the individual components of  
810 absorption for ocean color applications, *Journal of Geophysical Research*, 116, C02007,  
811 doi:10.1029/2009JC005594.  
812  
813 Mopper, K., D. Kieber, and A. Stubbins (2015), *Marine Photochemistry: Processes and Impacts*,  
814 in *Biogeochemistry of Marine Dissolved Organic Matter*, 2nd Edition, edited by D.A. Hansell  
815 and C.A. Carlson, Elsevier, San Diego, Calif.  
816  
817 Moran, M. A., and R. G. Zepp (1997), Role of Photoreactions in the Formation of Biologically  
818 Labile Compounds from Dissolved Organic Matter, *Limnology and Oceanography*, 42, 1307-  
819 1316.  
820  
821 Moran, M. A., W. M. Sheldon and R. G. Zepp (2000), Carbon loss and optical property changes  
822 during long-term photochemical and biological degradation of estuarine dissolved organic  
823 matter, *Limnol. Oceanogr.*, 45(6), 2000, 1254-1264.  
824  
825 Murphy, K.R., C.A. Stedmon, T.D. Waite, and G.M. Ruiz (2008), Distinguishing between  
826 terrestrial and autochthonous organic matter sources in marine environments using fluorescence  
827 spectroscopy, *Marine Chemistry*, 108, 40-58.  
828  
829 Murphy, K.R., K.D. Butler, R.G.M. Spencer, C.A. Stedmon, J.R. Boehme, G.R. Aiken (2010),  
830 Measurement of dissolved organic matter fluorescence in aquatic environments: An  
831 interlaboratory comparison, *Environmental, Science and Technology* 44, 9405-9412.  
832  
833 Murphy, K.R., A. Hambly, S. Singh, R.K. Henderson, A. Baker, R. Stuetz, and S.J. Khan (2011),  
834 Organic matter fluorescence in municipal water recycling schemes: toward a unified PARAFAC  
835 model. *Environ. Sci. Technol.*, 45, 2909-2916.  
836  
837 Murphy, K.R., R. Bro, and C.A. Stedmon (2014a), Chemometric analysis of organic matter  
838 fluorescence. in: Coble, P., Baker, A., Lead, J., Reynolds, D., Spencer, R. (Eds.). *Aquatic*  
839 *organic matter fluorescence*. Cambridge University Press, New York.  
840  
841 Murphy K.R., C.A. Stedmon, P. Wenig R. Bro (2014b), OpenFluor - A spectral database of auto-  
842 fluorescence by organic compounds in the environment. *Anal. Methods*, 6, 658-661.  
843 doi:10.1039/C3AY41935E  
844  
845 Nelson, N.B., and D.A. Siegel (2002) Chromophoric DOM in the open ocean, in *Biogeochemistry*  
846 *of Marine Dissolved Organic Matter*, edited by D.A. Hansell and C.A. Carlson, pp. 547-578,  
847 Elsevier, San Diego, Calif.  
848

849 Nicolaus, M., C. Katlein, J. Maslanik, and S. Hendricks (2012), Changes in Arctic sea ice result  
850 in increasing light transmittance and absorption, *Geophys. Res. Lett.*, 39, L24501,  
851 doi:10.1029/2012GL053738.  
852

853 Norman, L., D. N. Thomas, C. A. Stedmon, M. A. Granskog, S. Papadimitriou, R. Krapp, K. M.  
854 Meiners, D. Lannuzel, P. van der Merwe and G. S. Dieckmann (2011), The characteristics of  
855 chromophoric dissolved organic matter (CDOM) in Antarctic sea ice, *Deep-Sea Res. II*, 58,  
856 1075-1091, doi:10.1016/j.dsr2.2010.10.030.  
857

858 Obernosterer, I., B. Reitner and G. Herndl (1999), Contrasting effects of solar radiation on  
859 dissolved organic matter and its bioavailability to marine bacterioplankton, *Limnol. Oceanogr.*,  
860 44(7), 1645–1654.  
861

862 Obernosterer, I., R. Sempere and G. J. Herndl (2001) Ultraviolet radiation induces reversal of the  
863 bioavailability of DOM to marine bacterioplankton. *Aquatic Microbial Ecology*, 24(1), 61-68.  
864

865 Opsahl, S., R. Benner, and R. M. Amon (1999), Major flux of terrigenous dissolved organic  
866 matter through the Arctic Ocean, *Limnol. Oceanogr.*, 44,(8) 2017-2023,  
867 doi:10.4319/lo.1999.44.8.2017.  
868

869 Osburn, C.L., Morris, D.P., Thorn, K.A., Moeller, R.E., 2001. Chemical and optical changes in  
870 freshwater dissolved organic matter exposed to solar radiation. *Biogeochemistry*, 54(3): 251-278.  
871

872 Osburn, C. L., L. Retamal and W. F. Vincent (2009), Photoreactivity of chromophoric dissolved  
873 organic matter transported by the Mackenzie River to the Beaufort Sea, *Marine Chemistry*, 115,  
874 10-20, doi:10.1016/j.marchem.2009.05.003.  
875

876 Palm, S. P., S. T. Strey, J. Spinhirne and T. Markus (2010), Influence of Arctic sea ice extent on  
877 polar cloud fraction and vertical structure and implications for regional climate, *Journal of*  
878 *Geophysical Research*, 115, D21209, doi:10.1029/2010JD013900.  
879

880 Pegau, W. S. (2002), Inherent optical properties of the central Arctic surface waters, *Journal of*  
881 *Geophysical Research*, 107(C10), 8035, doi:10.1029/2000JC000382.  
882

883 Perovich, D. K., B. Light, H. Eicken, K. F. Jones, K. Runciman, and S. V. Nghiem (2007),  
884 Increasing solar heating of the Arctic Ocean and adjacent seas, 1979-2005: Attribution and role  
885 in the ice-albedo feedback, *Geophysical Research Letters*, 34, L19505,  
886 doi:10.1029/2007GL031480.  
887

888 Perovich, D., W. Meier, M. Tschudi, S. Gerland, J. Richter-Menge (2012), Sea Ice [In Arctic  
889 Report Card 2012], <http://www.arctic.noaa.gov/reportcard>.  
890

891 Peterson, B. J., R. M. Holmes, J. W. McClelland, C. J. Vorosmarty, I. A. Shiklomonov, R. B.  
892 Lammers, and S. Rahmstorf (2002), Increasing river discharge to the Arctic Ocean, *Science*, 298,  
893 2171-2173.  
894

895 Powers, L. C., and W. L. Miller (2015), Photochemical production of CO and CO<sub>2</sub> in the  
896 Northern Gulf of Mexico: Estimates and challenges for quantifying the impact of photochemistry  
897 on carbon cycles, *Marine Chemistry*, 171: 21-35.  
898  
899 Raymond, P.A. and R. G.M. Spencer, (2105), Riverine DOM, in *Biogeochemistry of Marine*  
900 *Dissolved Organic Matter* (Second Edition), edited by Dennis A. Hansell Craig A. Carlson,  
901 Academic Press, Boston, pp. 509-533, ISBN 9780124059405, [http://dx.doi.org/10.1016/B978-0-](http://dx.doi.org/10.1016/B978-0-12-405940-5.00011-X)  
902 [12-405940-5.00011-X](http://dx.doi.org/10.1016/B978-0-12-405940-5.00011-X).  
903  
904 Retamal, L., W. F. Vincent, C. Martineau and C. L. Osburn (2007), Comparison of the optical  
905 properties of dissolved organic matter in two river-influenced coastal regions of the Canadian  
906 Arctic, *Estuarine, Coastal and Shelf Science*, 72, 261-272, doi:10.1016/j.ecss.2006.10.022.  
907  
908 Retamal, L., Bonilla, S. et al. (2008), Optical gradients and phytoplankton production in the  
909 Mackenzie River and the coastal Beaufort Sea, *Polar Biology*, 31(3), 363-379.  
910  
911 Romanovsky, V. E., S. L. Smith and H. H. Christiansen (2010), Permafrost Thermal State in the  
912 Polar Northern Hemisphere during the International Polar Year 2007–2009: a Synthesis,  
913 *Permafrost and Periglac. Process.*, 21, 106–116, doi:10.1002/ppp.689.  
914  
915 Ruggaber, A., R. Dlugi and T. Nakajima (1994), Modelling radiation quantities and photolysis  
916 frequencies in the troposphere, *Journal of Atmospheric Chemistry*, 18(2), 171-210.  
917  
918 Schröder, D., D. L. Feltham, D. Flocco and M. Tsamados (2014), September Arctic sea-ice  
919 minimum predicted by spring melt-pond fraction, *Nature Climate Change*, 4, 353-357, doi:  
920 10.1038/nclimate2203.  
921  
922 Scully, N., and M. W. L. Miller (2000), Spatial and temporal dynamics of  
923 colored dissolved organic matter in the north water polynya, *Geophys.*  
924 *Res. Lett.*, 27, 1009–1011.  
925  
926 Seredynska-Sobecka, B., C.A. Stedmon, R. Boe-Hansen, C.K. Waul, and E. Arvin (2011),  
927 Monitoring organic loading to swimming pools by fluorescence excitation-emission matrix with  
928 parallel factor analysis (PARAFAC), *Water Research*, 45 (6), 2306-2314.  
929  
930 Serreze, M.C., M. M. Holland, and J. Stroeve (2007), Perspectives on the Arctic's shrinking sea  
931 ice cover, *Science*, 315, 133-1536, doi:10.1126/science.1139426.  
932  
933 Shutova, Y., A. Baker, J. Bridgeman and R.K. Henderson (2014), Spectroscopic characterisation  
934 of dissolved organic matter changes in drinking water treatment: from PARAFAC analysis to  
935 online monitoring wavelengths, *Water Research*, 54, 159-169.  
936  
937 Smith, L. C., T. M. Pavelsky, G. M. MacDonald, A. I. Shiklomanov and R. B. Lammers (2007),  
938 Rising minimum daily flows in northern Eurasian rivers: A growing influence of groundwater in  
939 the high-latitude hydrologic cycle, *Journal of Geophysical Research*, 112, G04S47,  
940 doi:10.1029/2006JG000327.

941  
 942 Solomon, S., R. W. Portmann, and D. W. J. Thomas (2007), Contrasts between Antarctic and  
 943 Arctic ozone depletion, *Proceedings of the National Academy of Sciences*, 104(2), 445-449.  
 944  
 945 Spencer, R.G.M., G.R. Aiken, K.P. Wickland, R.G. Striegl, P.J. Hernes (2008), Seasonal and  
 946 spatial variability in dissolved organic matter quantity and composition from the Yukon River  
 947 Basin, Alaska, *Global Biogeochemical Cycles*, 22, GB4002, doi:10.1029/ 2008GB003231.  
 948  
 949 Spencer, R. G. M., et al. (2009), Photochemical degradation of dissolved organic matter and  
 950 dissolved lignin phenols from the Congo River, *J. Geophys. Res.*, 114, G03010,  
 951 doi:10.1029/2009JG000968.  
 952  
 953 Spencer, R.G.M., K.D., Butler and G.R., Aiken (2012)' Dissolved organic carbon and  
 954 chromophoric dissolved organic matter properties of rivers in the U.S.A., *Journal of Geophysical*  
 955 *Research – Biogeosciences*, 117, G03001, doi:10.1029/2011JG001928.  
 956  
 957 Stedmon, C.A., and S. Markager (2001), The optics of chromophoric dissolved organic matter  
 958 (CDOM) in the Greenland Sea: An algorithm for differentiation between marine and terrestrially  
 959 derived organic matter, *Limnol. Oceanogr.*, 46, 2087–2093.  
 960  
 961 Stedmon , C. A., S. Markager, R. Bro (2003), Tracing dissolved organic matter in aquatic  
 962 environments using a new approach to fluorescence spectroscopy, *Mar. Chem.*, 82, 239-254,  
 963 doi:10.1016/S0304-4203(03)00072-0.  
 964  
 965 Stedmon, C.A. and S. Markager (2005a), Resolving the variability of dissolved organic matter  
 966 fluorescence in a temperate estuary and its catchment using PARAFAC analysis, *Limnology and*  
 967 *Oceanography*, 50, 686–697.  
 968  
 969 Stedmon, C.A. and S. Markager (2005b), Tracing the production and degradation of  
 970 autochthonous fractions of dissolved organic matter using fluorescence analysis, *Limnology and*  
 971 *Oceanography*, 50 (5), 1415–1426.  
 972  
 973 Stedmon, C.A., S. Markager, L. Tranvik, L. Kronberg, T. Slätis and W. Martinsen (2007a),  
 974 Photochemical production of ammonium and transformation of dissolved organic matter in the  
 975 Baltic Sea, *Mar. Chem.*, 104, 227-240, doi:10.1016/j.marchem.2006.11.005.  
 976  
 977 Stedmon, C.A., D.N. Thomas, M. Granskog, H. Kaartokallio, S. Papadimitriou, and H. Kuosa  
 978 (2007b), Characteristics of Dissolved Organic Matter in Baltic Coastal Sea Ice: Allochthonous or  
 979 Autochthonous Origins?, *Environ. Sci. Technol.*, 41, 7273-7279.  
 980  
 981 Stedmon, C.A., and R. Bro (2008), Characterizing dissolved organic matter fluorescence with  
 982 parallel factor analysis: A tutorial, *Limnol. Oceanogr. Methods*, 6, 572-579.  
 983  
 984 Stedmon, C.A., R.M.W. Amon, A.J. Rinehart, S.A. Walker (2011), The supply and  
 985 characteristics of colored dissolved organic matter (CDOM) in the Arctic Ocean: Pan Arctic  
 986 trends and differences, *Mar. Chem.*, 124, 108-118, doi:10.1016/j.marchem.2010.12.007.

987  
988 Stroeve, J., M. M. Holland, W. Meier, T. Scambos, and M. Serreze (2007), Arctic sea ice  
989 decline: Faster than forecast, *Geophys. Res. Lett.*, 34, L09501, doi:10.1029/2007GL029703.  
990  
991 Stubbins, A., V. Hubbard, G. Uher, C. S. Law, R. C. Upstill-Goddard, G. R. Aiken, and K.  
992 Mopper (2008), Relating Carbon Monoxide Photoproduction to Dissolved Organic Matter  
993 Functionality, *Environ. Sci. Technol.*, 42, 3271-3276, doi.  
994  
995 Stubbins, A., R. G. M. Spencer, H. Chen, P. G. Hatcher, K. Mopper, P. J. Hernes, V. L.  
996 Mwamba, A. M. Mangangu, J. N. Wabakanghanzi, and J. Six (2010), Illuminated darkness:  
997 molecular signatures of Congo River dissolved organic matter and its photochemical alteration as  
998 revealed by ultrahigh precision mass spectrometry, *Limnol. Oceanogr.*, 55, 1467-1477,  
999 doi:10.4319/lm.2010.55.4.1467.  
1000  
1001 Stubbins, A., and T. Dittmar (2012), Low volume quantification of dissolved organic carbon and  
1002 dissolved nitrogen, *Limnol. Oceanogr. Methods*, 10, 347-352, doi:10.4319/lom.2012.10.347.  
1003  
1004 Stubbins, A., J.-F. Lapierre, M. Berggren, Y. Prairie, T. Dittmar, and P. del Giorgio, (2014),  
1005 What's in an EEM? Molecular signatures associated with dissolved organic fluorescence in  
1006 boreal Canada, *Environmental Science and Technology*, doi: 10.1021/es502086e.  
1007  
1008 Tanaka, K., K. Kuma, K. Hamasaki, and Y. Yamashita (2014) Accumulation of humic-like  
1009 fluorescent dissolved organic matter in the Japan Sea, *Scientific Reports*, 4, 5292.  
1010  
1011 Tranvik, L. and S. Kokalj (1998) Decreased biodegradability of algal DOC due to interactive  
1012 effects of UV radiation and humic matter, *Aquat. Microb. Ecol.*, 14, 301-307.  
1013  
1014 Tranvik, L. J., H. Olofsson and S. Bertilsson (1999) Photochemical Effects on Bacterial  
1015 Degradation of Dissolved Organic Matter in Lake Water, in *Microbial Biosystems: New*  
1016 *Frontiers Proceedings of the 8th International Symposium on Microbial Ecology*  
1017 edited by C. R. Bell, M. Brylinsky and P. Johnson-Green, Atlantic Canada Society for Microbial  
1018 Ecology, Halifax, Canada.  
1019  
1020 Tranvik, L. J., and S. Bertilsson (2001), Contrasting effects of solar UV radiation on  
1021 dissolved organic sources for bacterial growth, *Ecol. Lett.*, 4, doi: 10.1046/j.1461-  
1022 0248.2001.00245.x  
1023  
1024 Vahatalo, A. V. and R. G. Wetzel (2004) Photochemical and microbial decomposition of  
1025 chromophoric dissolved organic matter during long (months-years) exposures, *Marine*  
1026 *Chemistry*, 89, 313- 326, doi:10.1016/j.marchem.2004.03.010.  
1027  
1028 Vavrus, S., M. M. Holland, and D. A. Bailey (2010), Changes in Arctic clouds during intervals  
1029 of rapid sea ice loss, *Climate Dynamics*, 36, 1475-1489, doi:10.1007/s00382-010-0816-0, 2010.  
1030  
1031 Walker, S. A., R. M. W. Amon, C. Stedmon, S. Duan, and P. Louchouart (2009), The use of  
1032 PARAFAC modeling to trace terrestrial dissolved organic matter and fingerprint water masses in

1033 coastal Canadian Arctic surface waters, *J. Geophys. Res.*, 114, G00F06,  
1034 doi:10.1029/2009JG000990.  
1035  
1036 Wang, X. and J. R. Key (2005), Arctic Surface, Cloud, and Radiation Properties Based on the  
1037 AVHRR Polar Pathfinder Dataset. Part II: Recent Trends, *Journal of Climate*, 18, 2575-2593.  
1038  
1039 Wang, M. and J. E. Overland (2009), A sea ice free summer Arctic within 30 years?, *Geophys.*  
1040 *Res. Lett.*, 36, L07502, doi:10.1029/2009GL037820.  
1041  
1042 Williamson, C.E., Neale, P.J et al. (2001), Beneficial and detrimental effects of UV on aquatic  
1043 organisms: Implications of spectral variation, *Ecological Applications*, 11(6), 1843–1857.  
1044  
1045 Yamashita, Y., R. M. Cory, J. Nishioka, K. Kuma, E. Tanoue, R. Jaffé (2010a), Fluorescence  
1046 characteristics of dissolved organic matter in the deep waters of the Okhotsk Sea and the  
1047 northwestern North Pacific Ocean, *Deep-sea Research II*, 57, 1478–1485.  
1048  
1049 Yamashita, Y., L.J. Scinto, N. Maie, and R. Jaffe (2010b), Dissolved Organic Matter  
1050 Characteristics Across a Subtropical Wetland Landscape: Application of Optical Properties in  
1051 the Assessment of Environmental Dynamics, *Ecosystems*, 13, 1006-1019.  
1052  
1053 Yamashita, Y., B.D. Kloeppel, J. Knoepp, G.L. Zausen, and R. Jaffé (2011), Effects of  
1054 Watershed History on Dissolved Organic Matter Characteristics in Headwater Streams,  
1055 *Ecosystems*, 14, 1110-1122.  
1056  
1057 Yamashita, Y., J.N. Boyer, and R. Jaffe (2013), Evaluating the distribution of terrestrial  
1058 dissolved organic matter in a complex coastal ecosystem using fluorescence spectroscopy,  
1059 *Continental Shelf Research*, 66, 136-144.

1060 **Table 1.** Initial, final and percent remaining after 72 hour irradiation experiments for  $a_{254}$  and DOC for all experiments. Bold numbers  
1061 indicate statistically significant changes ( $p < 0.05$ ). *NA* indicates that the value is unavailable owing to sample loss.  
1062

Sample Name	Collection Date	Salinity		$a_{254}$ ( $\text{m}^{-1}$ )	DOC ( $\mu\text{M}$ )
ICM <sub>1</sub>	7/4/2011	17.4	Initial	$1.41 \pm 0.03$	$67.8 \pm 0.8$
			Final	$0.66 \pm 0.02$	$65.8 \pm 1.2$
			% Remaining	<b>47.1</b>	97.0
ICM <sub>2</sub>	7/5/2011	30.7	Initial	$1.55 \pm 0.01$	$71.9 \pm 2.8$
			Final	$0.72 \pm 0.02$	$68.7 \pm 1.2$
			% Remaining	<b>46.5</b>	95.6
ICM <sub>3</sub>	7/6/2011	29.7	Initial	$1.37 \pm 0.04$	$61.1 \pm 2.5$
			Final	$0.62 \pm 0.01$	$60.6 \pm 0.4$
			% Remaining	<b>44.8</b>	99.1
ICM <sub>4</sub>	7/10/2011	22.4	Initial	$1.36 \pm 0.02$	$53.6 \pm 3.9$
			Final	$0.63 \pm 0.01$	$53.7 \pm 2.4$
			% Remaining	<b>46.5</b>	100.2
ICM <sub>5</sub>	7/13/2011	24.9	Initial	$2.17 \pm 0.01$	$71.1 \pm 0.1$
			Final	$0.98 \pm 0.03$	$68.0 \pm 1.3$
			% Remaining	<b>44.9</b>	95.6
ICM <sub>6</sub>	07/19/11	20.8	Initial	$2.01 \pm 0.03$	$69.4 \pm 0.6$
			Final	$0.98 \pm 0.02$	$67.0 \pm 0.8$
			% Remaining	<b>48.7</b>	<b>96.6</b>
IFM <sub>1</sub>	7/3/2011	31.7	Initial	$1.49 \pm 0.04$	<i>NA</i>
			Final	$0.77 \pm 0.04$	
			% Remaining	<b>51.8</b>	
IFM <sub>2</sub>	7/9/2011	31.1	Initial	$1.89 \pm 0.01$	$71.2 \pm 0.9$
			Final	$0.83 \pm 0.02$	$71.9 \pm 1.1$
			% Remaining	<b>44.0</b>	100.9
IFT <sub>1</sub>	07/21/11	28.0	Initial	$5.79 \pm 0.03$	$105.4 \pm 1.4$
			Final	$2.17 \pm 0.05$	$97.1 \pm 2.5$
			% Remaining	<b>37.4</b>	<b>92.1</b>
IFT <sub>2</sub>	07/20/11	24.4	Initial	$3.12 \pm 0.00$	$80.5 \pm 1.2$
			Final	$1.48 \pm 0.06$	$74.7 \pm 0.6$
			% Remaining	<b>47.4</b>	<b>92.8</b>

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**Table 2.** Average initial and final values for the 72 hour irradiation experiments for  $a_{254}$ ,  $a_{365}$  and  $a_{440}$  for ice-covered marine (ICM), ice-free marine (IFM), and ice-free terrestrial (IFT) samples.

		$a_{254}$ ( $\text{m}^{-1}$ )	$a_{365}$ ( $\text{m}^{-1}$ )	$a_{440}$ ( $\text{m}^{-1}$ )
ICM <sub>mean</sub>	Initial	$1.62 \pm 0.32$	$0.14 \pm 0.02$	$0.04 \pm 0.01$
	Final	$0.76 \pm 0.16$	$0.06 \pm 0.01$	$0.02 \pm 0.01$
IFM <sub>mean</sub>	Initial	$1.69 \pm 0.22$	$0.17 \pm 0.03$	$0.06 \pm 0.01$
	Final	$0.80 \pm 0.05$	$0.08 \pm 0.02$	$0.04 \pm 0.01$
IFT <sub>mean</sub>	Initial	$4.46 \pm 1.46$	$0.50 \pm 0.29$	$0.13 \pm 0.08$
	Final	$1.82 \pm 0.38$	$0.18 \pm 0.08$	$0.05 \pm 0.03$

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**Table 3.** Three parameter exponential decay models for  $a_{254}$  vs. cumulative UV light absorbed for all experiments.

	ICM <sub>1</sub>	ICM <sub>2</sub>	ICM <sub>3</sub>	ICM <sub>4</sub>	ICM <sub>5</sub>	ICM <sub>6</sub>	IFM <sub>1</sub>	IFM <sub>2</sub>	IFT <sub>1</sub>	IFT <sub>2</sub>
$z_0$	0.78	0.90	0.80	0.80	1.28	1.08	0.73	1.17	4.21	1.77
$K$	1.64	2.25	1.94	1.74	1.41	1.54	1.96	0.82	0.25	0.81
$C_\infty$	0.63	0.71	0.59	0.57	0.89	0.91	0.76	0.68	1.63	1.36
$C(\theta)$	1.42	1.60	1.39	1.36	2.17	1.99	1.49	1.86	5.84	3.13
$R^2$	0.96	0.98	0.97	0.97	0.96	0.96	0.96	0.94	0.98	0.98

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**Table 4.** Initial, final and percent remaining after 72 hour irradiation experiments for the spectral slope parameters  $S_{275-295}$ ,  $S_{350-400}$ , and  $S_R$  for all experiments. Bold numbers indicate statistically significant changes ( $p < 0.05$ ).

		ICM <sub>1</sub>	ICM <sub>2</sub>	ICM <sub>3</sub>	ICM <sub>4</sub>	ICM <sub>5</sub>	ICM <sub>6</sub>	IFM <sub>1</sub>	IFM <sub>2</sub>	IFT <sub>1</sub>	IFT <sub>2</sub>
$S_{275-295}$	Initial	0.0241	0.0266	0.0264	0.0299	0.0311	0.0337	0.0255	0.0260	0.0203	0.0284
	Final	0.0233	0.0238	0.0238	0.0257	0.0259	0.0255	0.0197	0.0249	0.0202	0.0242
	Difference	<b>-0.0008</b>	<b>-0.0028</b>	<b>-0.0026</b>	<b>-0.0043</b>	<b>-0.0052</b>	<b>-0.0081</b>	<b>-0.0057</b>	<b>-0.0011</b>	-0.0001	<b>-0.0041</b>
$S_{350-400}$	Initial	0.0180	0.0170	0.0185	0.0217	0.0161	0.0197	0.0148	0.0170	0.0184	0.0202
	Final	0.0136	0.0158	0.0116	0.0139	0.0181	0.0188	0.0121	0.0109	0.0171	0.0198
	Difference	<b>-0.0044</b>	<b>-0.0012</b>	<b>-0.0069</b>	<b>-0.0078</b>	0.0021	<b>-0.0009</b>	<b>-0.0026</b>	<b>-0.0060</b>	<b>-0.0013</b>	<b>-0.0004</b>
$S_R$	Initial	1.3371	1.5692	1.4257	1.3830	1.9370	1.7120	1.7253	1.5275	1.1029	1.4046
	Final	1.7154	1.5036	2.0480	1.8492	1.4288	1.3581	1.6264	2.2728	1.1793	1.2241
	Difference	<b>0.3783</b>	-0.0656	<b>0.6223</b>	<b>0.4662</b>	-0.5081	<b>-0.3539</b>	-0.0989	<b>0.7452</b>	<b>0.0764</b>	-0.1805

1098 **Table 5.** Spectral characteristics of the six components identified by PARAFAC compared to previously identified components.  
1099 Secondary maxima are shown in parentheses. Component matches (>0.95 tucker congruent coefficient) identified using the OpenFluor  
1100 database (<http://www.openfluor.org>, Murphy et al. [2014b]) are in bold. Matches not in bold were based on visual inspection.  
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Components	Excitation Maxima (nm)	Emission Maxima (nm)	Other Studies	Description
1	<250 (330)	468	<b>C1 (Jørgensen et al., 2011)</b> <b>C1 (Kowalczuk et al., 2009)</b> <b>G1 (Murphy et al., 2011)</b> <b>C3 (Seredynska-Sobecka et al, 2011)</b> <b>C1 (Shutova et al., 2014)</b> <b>C1 (Stedmon and Markager, 2005b)</b> <b>C1 (Stedmon et al., 2007a)</b> <b>C1 (Tanaka et al., 2014)</b> <b>C1 (Yamashita et al., 2010b)</b>	humic-like, terrestrial
2	255 (280)	392	C3 (Stedmon and Markager, 2005a) C2 and C3 (Stedmon and Markager, 2005b)	humic-like, terrestrial
3	280	340	<b>C7 (Murphy et al., 2008)</b> <b>C5 (Murphy et al. 2014a)</b>	Amino acids, free or bound in proteins
4	270	320	<b>C7 (Graeber et al., 2012)</b> <b>C4 (Yamashita et al., 2011)</b> <b>C3 (Yamashita et al., 2013)</b>	Protein- and tyrosine-like, positively related to bioavailability
5	<250	352	<b>C6 (Jørgensen et al., 2011)</b>	protein-like, resembles free tryptophan
6	<250 (325)	392	C4 (Jørgensen et al., 2011) C6 (Stedmon and Markager, 2005a) C2 (Stedmon et al., 2007a) C3 (Stedmon et al., 2007b) C2 (Yamashita et al., 2010a)	humic-like, marine also common in wastewater and agricultural catchments

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**Table 6.** Fluorescence intensities (R.U.) of individual components identified using PARAFAC at the beginning and end of the 72 h irradiation. Percent of total fluorescence intensity is shown in parentheses.

Components	ICM <sub>1</sub>	ICM <sub>2</sub>	ICM <sub>3</sub>	ICM <sub>4</sub>	ICM <sub>5</sub>	ICM <sub>6</sub>	IFM <sub>1</sub>	IFM <sub>2</sub>	IFT <sub>1</sub>	IFT <sub>2</sub>
<b>1</b>	0.022 (13)	0.026 (21)	0.022 (16)	0.019 (17)	0.027 (20)	0.021 (16)	0.022 (6)	0.036 (25)	0.136 (40)	0.049 (29)
	0.017 (27)	0.019 (23)	0.016 (26)	0.018 (20)	0.025 (22)	0.024 (43)	0.017 (17)	0.021 (17)	0.060 (56)	0.044 (48)
<b>2</b>	0.008 (5)	0.018 (15)	0.008 (6)	0.013 (11)	0.019 (14)	0.029 (22)	0.001 (0)	0.015 (10)	0.063 (18)	0.029 (17)
	0.004 (7)	0.005 (6)	0.004 (6)	0.005 (5)	0.008 (7)	0.008 (15)	0.000 (0)	0.003 (2)	0.015 (14)	0.014 (15)
<b>3</b>	0.037 (22)	0.020 (16)	0.031 (22)	0.016 (15)	0.020 (14)	0.000 (0)	0.085 (22)	0.026 (18)	0.030 (9)	0.011 (7)
	0.018 (28)	0.023 (28)	0.007 (12)	0.007 (8)	0.021 (19)	0.002 (4)	0.031 (31)	0.031 (25)	0.007 (7)	0.005 (6)
<b>4</b>	0.036 (21)	0.032 (26)	0.032 (23)	0.028 (25)	0.030 (21)	0.074 (56)	0.075 (20)	0.029 (20)	0.035 (10)	0.030 (18)
	0.017 (25)	0.017 (20)	0.014 (24)	0.023 (26)	0.006 (5)	0.007 (12)	0.034 (34)	0.026 (21)	0.004 (4)	0.006 (7)
<b>5</b>	0.054 (32)	0.014 (11)	0.037 (27)	0.028 (25)	0.031 (23)	0.000 (0)	0.180 (48)	0.021 (15)	0.037 (11)	0.032 (19)
	0.001 (1)	0.010 (12)	0.013 (21)	0.027 (31)	0.039 (35)	0.003 (5)	0.010 (10)	0.031 (26)	0.001 (0)	0.005 (6)
<b>6</b>	0.011 (7)	0.014 (11)	0.010 (7)	0.007 (6)	0.012 (8)	0.014 (7)	0.014 (4)	0.016 (11)	0.044 (13)	0.018 (11)
	0.008 (12)	0.009 (11)	0.007 (11)	0.008 (9)	0.013 (11)	0.008 (20)	0.008 (9)	0.011 (9)	0.020 (19)	0.018 (19)

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1121 | **Figure Captions**

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1123 **Figure 1.** Station locations for under-ice marine samples (white), ice-free marine samples (blue)  
1124 and terrestrially-influenced samples (red). The location of the ice edge on the first day of  
1125 sampling (July 3, 2011) and the last day of sampling (July 21, 2011) are indicated by the purple  
1126 and blue lines respectively.

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1128 **Figure 2.** Loss of CDOM absorbance with irradiation for triplicate samples from experiments (a)  
1129 ICM<sub>1</sub>, (b) IFM<sub>1</sub> and (c) IFT<sub>1</sub>. CDOM absorbance is shown at the start of the experiment ( $t_0$ ) and  
1130 after 4, 12, 24 and 72 hours ( $t_4$ ,  $t_{12}$ ,  $t_{24}$  and  $t_{72}$ ).

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1132 **Figure 3.** Loss of CDOM absorbance at 254 nm as a function of (a) time and (b) the amount of  
1133 cumulative UV light (280–400 nm) absorption at each time interval for all experiments.

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1135 **Figure 4.** Relationship between the initial CDOM absorption coefficient at 254 nm ( $C(0)$ ) and  
1136 the photoreactive component ( $z_0$ ) from the three parameter exponential decay model.

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1138 **Figure 5.** Changes in the spectral slope  $S_{275-295}$  (a) and  $S_{350-400}$  (b) at each time interval for all  
1139 experiments.

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1141 **Figure 6.** EEMs from a representative under-ice sample (ICM<sub>1</sub>), ice-free marine sample (IFM<sub>1</sub>),  
1142 terrestrially-influenced sample (IFT<sub>1</sub>), and a sample that shows protein-like production (IFT<sub>2</sub>).

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1144 **Figure 7.** The six-components for the PARAFAC model. The top two rows show the excitation  
1145 (red lines) and emission (blue lines) loadings for each component and the bottom two rows show  
1146 the individual components.

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1148 **Figure 8.** Variability in fluorescence for the six PARAFAC components for all experiments over  
1149 the 72-hour irradiation experiments.

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1151 **Figure 9.** (a) Changes in CDOM absorbance at 254 nm as a function of the change in DOC after  
1152 72 hours for all experiments. (b) Percent loss in CDOM absorbance at 254 nm after 72 hours as a  
1153 function of the percent loss in DOC for all experiments.

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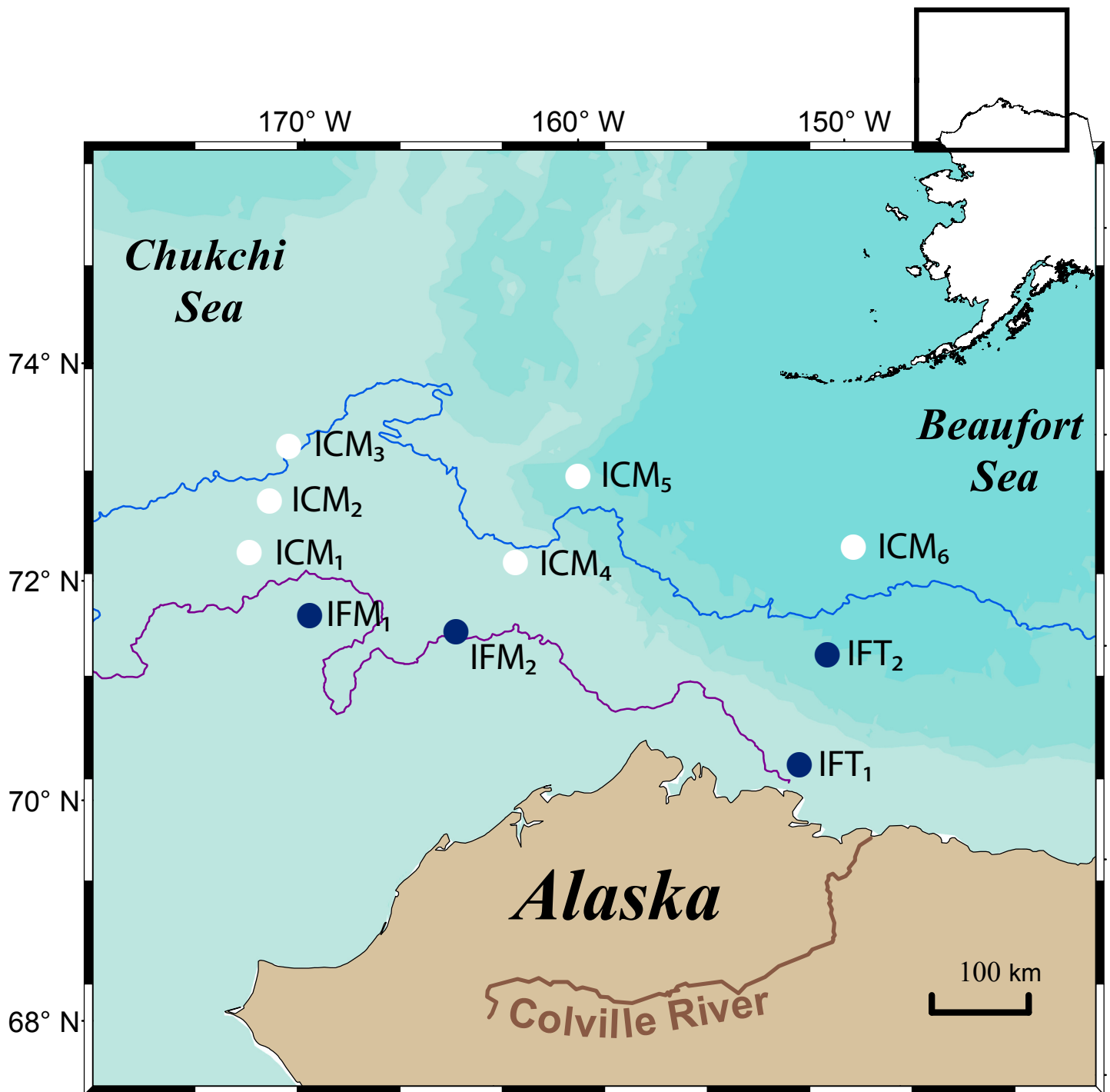
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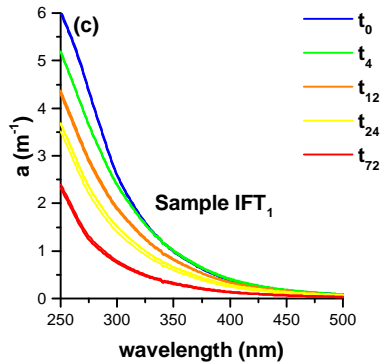
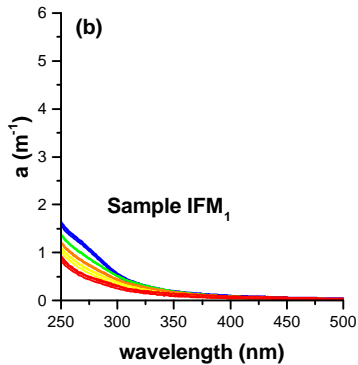
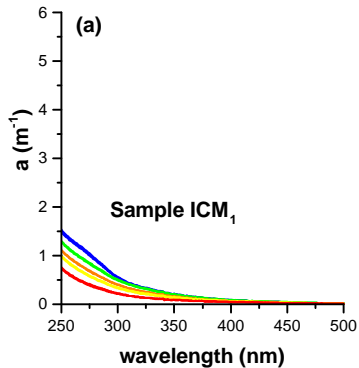
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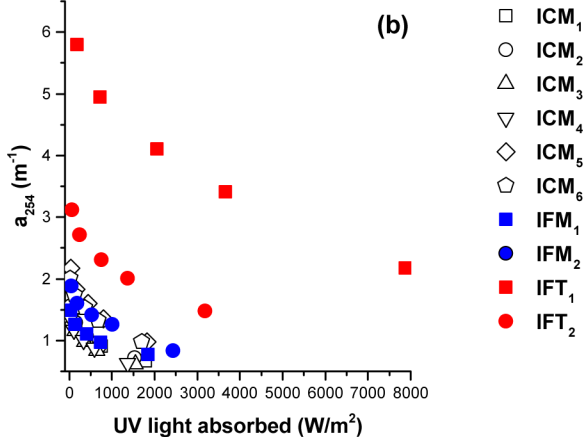
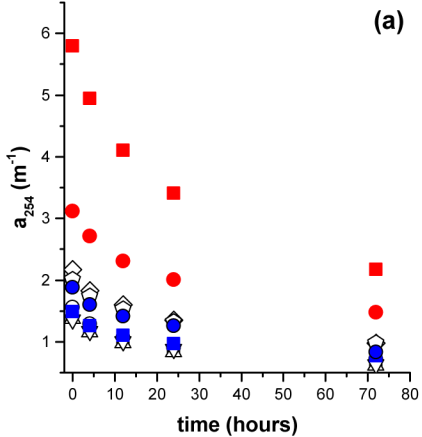
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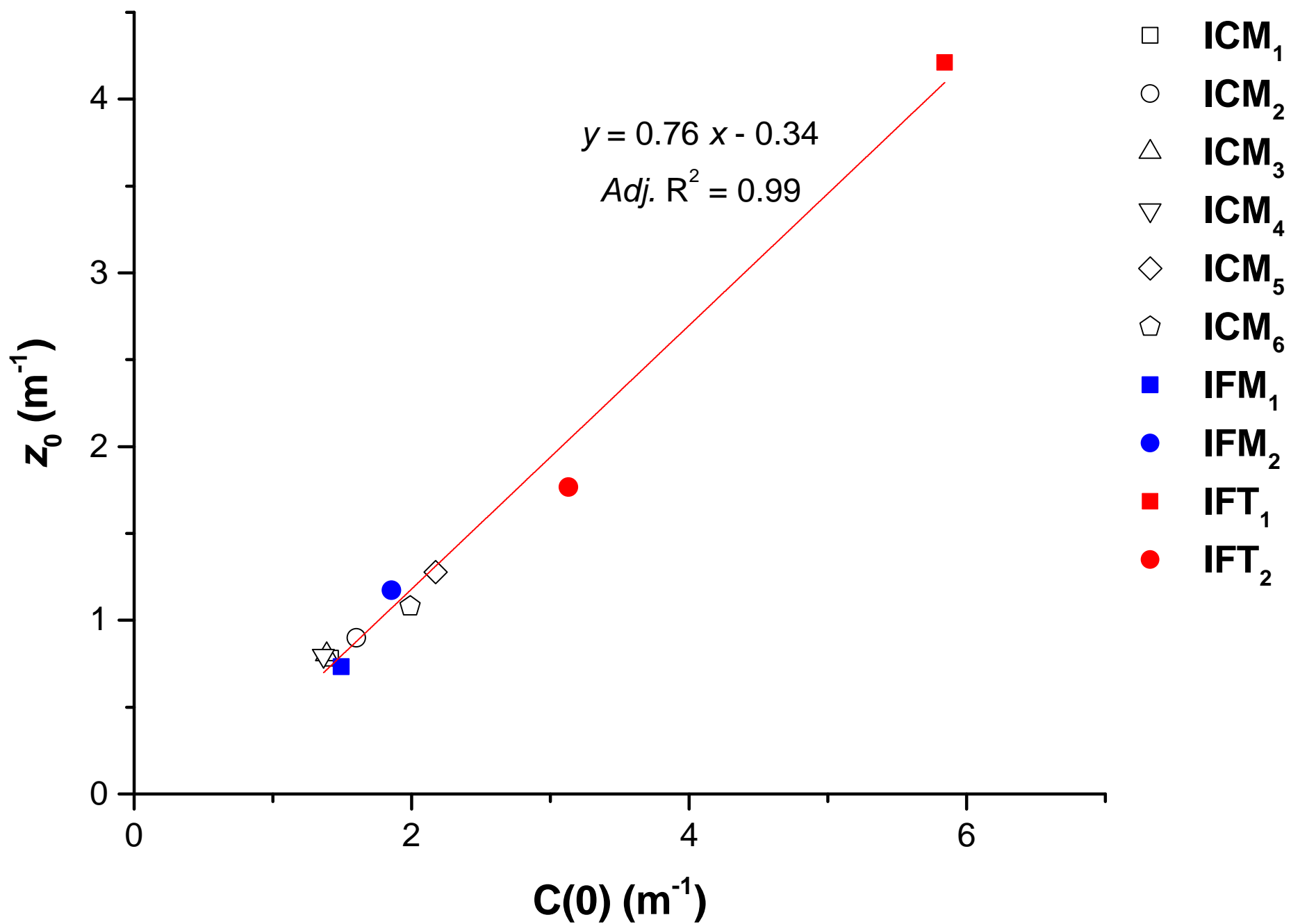
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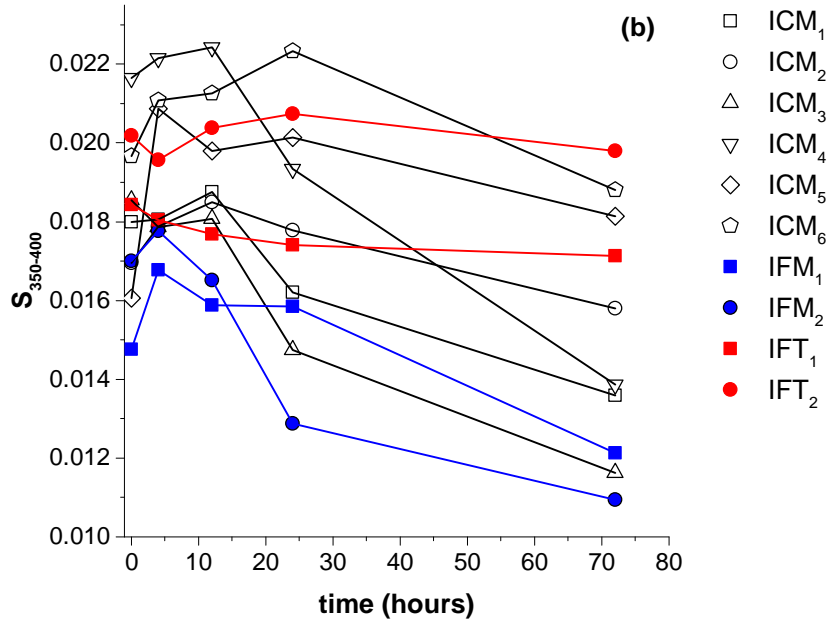
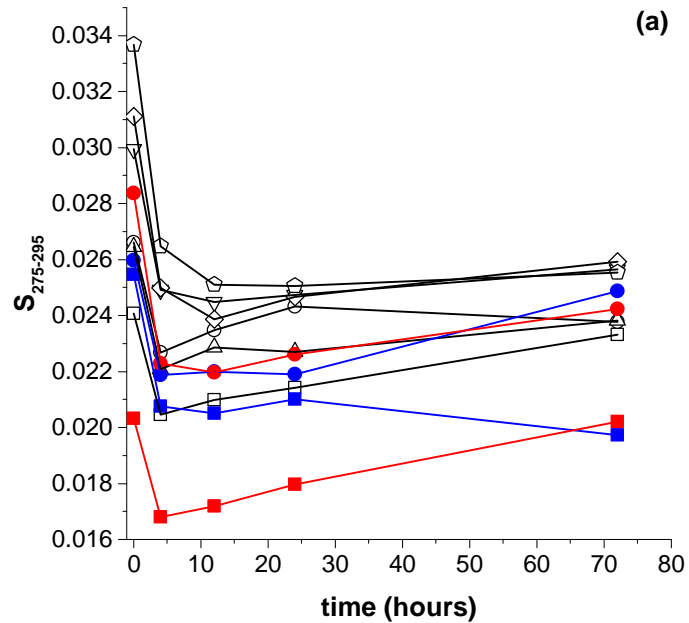


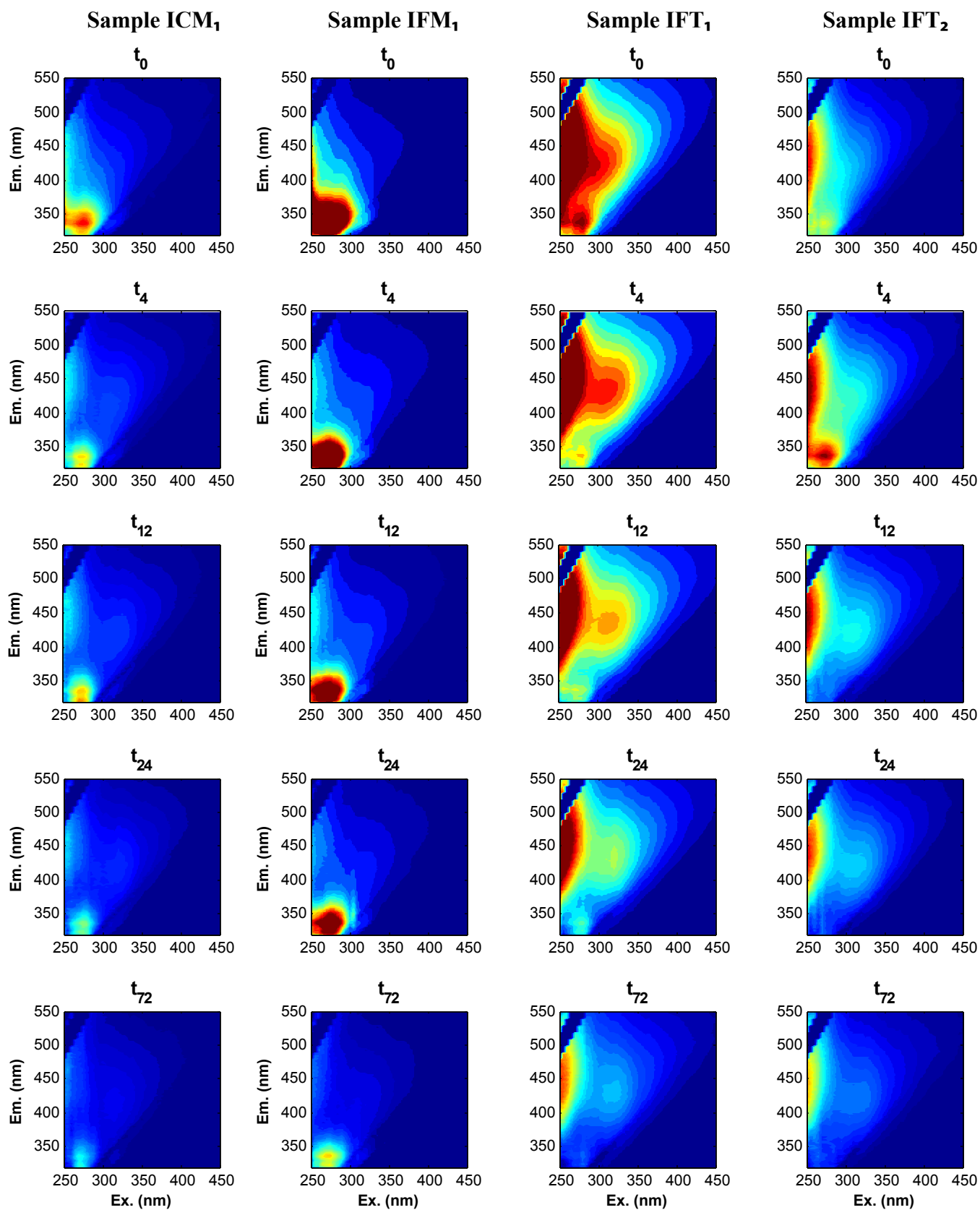


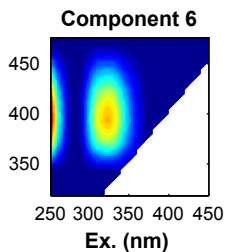
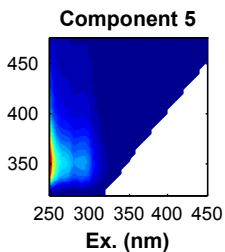
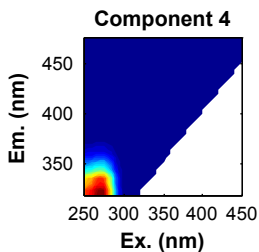
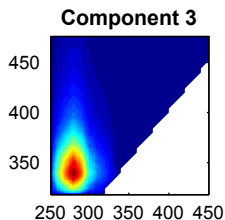
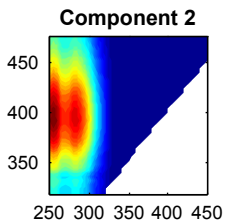
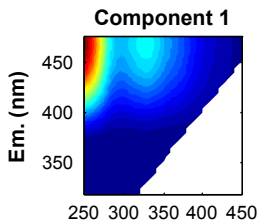
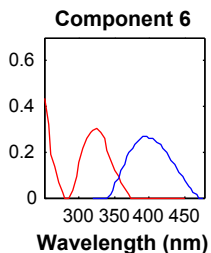
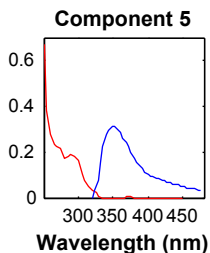
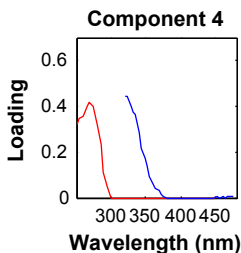
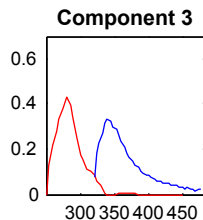
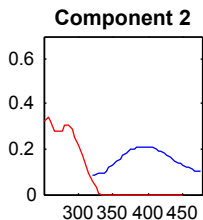
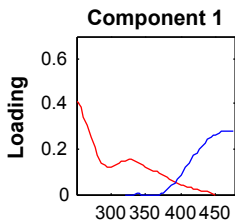


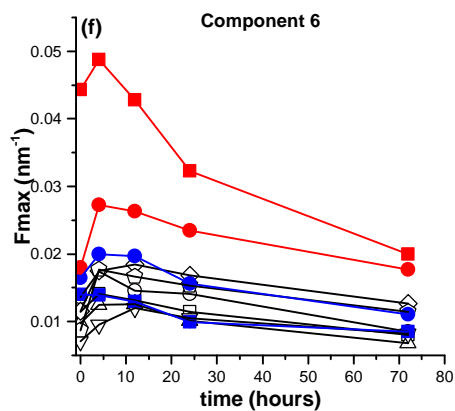
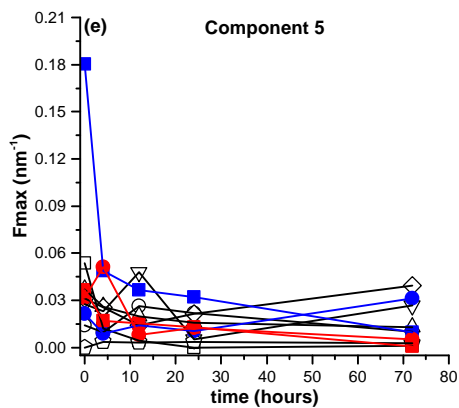
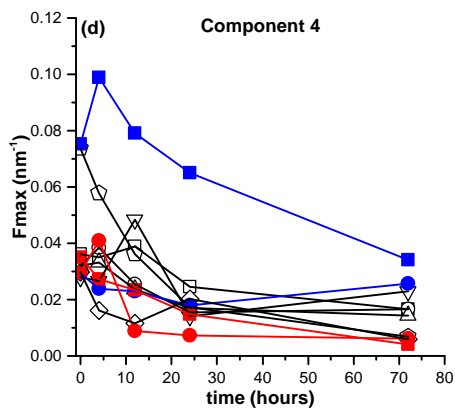
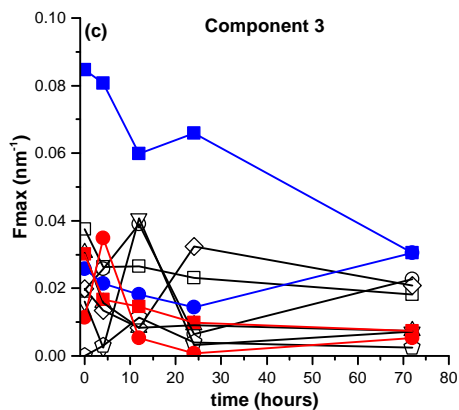
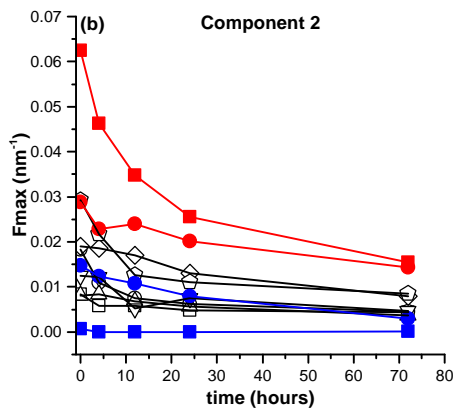
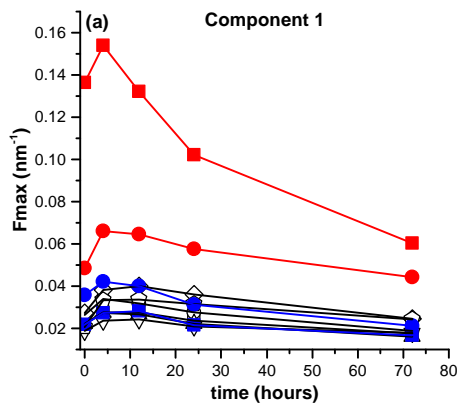












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- IFM<sub>1</sub>
- IFM<sub>2</sub>
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